



Citation: Alía R, Notivol E, Climent J, Pérez F, Barba D, Majada J, et al. (2022) Local seed sourcing for sustainable forestry. PLoS ONE 17(12): e0278866. https://doi.org/10.1371/journal.pone.0278866

Editor: Pankaj Bhardwaj, Central University of Punjab, INDIA

Received: August 10, 2022

Accepted: November 26, 2022

Published: December 14, 2022

Peer Review History: PLOS recognizes the benefits of transparency in the peer review process; therefore, we enable the publication of all of the content of peer review and author responses alongside final, published articles. The editorial history of this article is available here: https://doi.org/10.1371/journal.pone.0278866

Copyright: © 2022 Alía et al. This is an open access article distributed under the terms of the Creative Commons Attribution License, which permits unrestricted use, distribution, and reproduction in any medium, provided the original author and source are credited.

Data Availability Statement: Data are available from the Zenodo repository (DOI: 10.5281/zenodo. 7157589).

Funding: This research was supported by the Project AEG 17-048 established in the frame of the measure 15.2" support to the conservation and use

RESEARCH ARTICLE

Local seed sourcing for sustainable forestry

Ricardo Alía^{1*}, Eduardo Notivol², José Climent¹, Felipe Pérez³, Diana Barba¹, Juan Majada⁴, José Manuel García del Barrio¹

- 1 Department of Ecology and Forest Genetics, Institute of Forest Sciences, INIA-CSIC, Madrid, Spain,
- 2 Department of Environment, Agricultural and Forest Systems, CITA, Zaragoza, Spain, 3 Directorate General of Biodiversity, Forest and Desertification, MITECO, Madrid, Spain, 4 CETEMAS, Carbayin, Spain
- * alia@inia.csic.es

Abstract

Seed sourcing strategies are the basis for identifying genetic material meeting the requirements of future climatic conditions and social demands. Specifically, local seed sourcing has been extensively promoted, based on the expected adaptation of the populations to local conditions, but there are some limitations for the application. We analyzed Strict-sense local and Wide-sense local (based on climatic similarity) seed sourcing strategies. We determined species and genetic pools based on these strategies for 40 species and deployment zones in Spain. We also obtained the total number of seed sources and stands for these species in the EU countries. We analyzed the richness of the pools, the relationship with variables related to the use of the species in afforestation, and the availability of seed production areas approved for the production of reproductive material destined to be marketed. This study confirms the existence of extensive species and genetic local pools. Also, that the importance of these pools differs for different species, limitations being derived from the use of forest reproductive material and the existence of approved basic materials. Strategies derived from local seed sourcing approaches are the basis for the use of forest reproductive material because a large number of the species in the area considered in the study are under regulation. However, despite the extensive work done to approve basic materials, limitations based on the availability of seed production areas to provide local material for sustainable forestry are found in those species. Considering a Wide-sense local seed sourcing strategy we provide alternative pools in order to meet social demands under the actual regulations on marketing of reproductive materials.

Introduction

Sustainable forest management aims at maintaining the biodiversity, productivity, regeneration capacity, vitality and potential of forests (Resolution H1, Forest Europe 1993), taking into account the economic value of the ecosystem services they provide [1], and the urgent need to increase forest resilience [2]. In this context, both artificial and natural regeneration play essential roles in ensuring resilience [3], long-term population persistence [4] and the restoration of ecosystem functionality [5, 6].

of forest genetic resources" and under Regulation (EU) No 1305/2013 of the European Parliament and of the Council of 17 December 2013 on support for rural development by the European Agricultural Fund for Rural Development (EAFRD) with 75% co-financing (FP). This study was funded by the European Union Horizon 2020 research and innovation programme under grant agreement No 773383 (B4EST project) (RA), and the Ministry of Science (RTI2018-094691-B-C32) (JC). The funders had no role in study design, data collection and analysis, decision to publish, or preparation of the manuscript.

Competing interests: The authors have declared that no competing interests exist.

While natural regeneration is always based on local genetic resources, artificial regeneration often makes use of the deliberate transfer of genetic resources from elsewhere. In the last case, seed sourcing involves matching the reproductive material collection area (Seed Production Area (SPA, [7]), region of provenance or seed zone) and the area where the material will be used (deployment zone) [8].

The origin of the planting stock is a major concern in forestry since it influences (as shown by extensive provenance research [9]) the existing genetic resources [10] and the future performance and adaptability [11] of populations. The election of a local provenance (*local seed sourcing*) is based on the expectation that populations are locally adapted [12–14], as origin also influences different traits related to the stability, adaptation, resistance, productivity and diversity of the planting stock [14, 15]. Local seed sourcing is the preferred, most widespread approach [16] in biodiversity conservation [17]. It is recommended as a general principle in Forest Europe and is followed by many countries [3]. It can be interpreted as *strict-sense* when the material used in the deployment location was also collected there, or as *wide-sense* when the material used in the deployment location was collected from locations with similar climate [16, 18].

Other strategies have been suggested under climate change scenarios, such as climate-predictive, composite, admixture, climate-adjusted, and assisted migration [16, 19], which require information on the performance of the planting stock under different environments [20–27]. A given seed-sourcing strategy makes it possible to define a species pool, i.e. the set of species that can potentially inhabit a site due to suitable local ecological conditions [28], and a genetic pool within those species [29], i.e. the set of genetic materials suitable for the site.

Although these pools can be defined only theoretically, their use for specific purposes (e.g., for forestry), requires that SPA should meet the regulations (e.g. European Union, OCDE, USA schemes [30–32]) to be approved for the production of reproductive materials (i.e. seeds, fruits, plants or part of plants). These regulations define different basic materials from which they can be obtained (e.g. seed-sources, stands, seed orchards, parents of families, clones and mixture of clones) and different categories of reproductive materials (e.g. source-identified, selected, qualified and tested). They also define basic units for the marketing of source-identified and selected categories of reproductive materials (regions of provenance or seed zones, depending on the scheme). Among all the types of basic materials, SPA for producing sourceidentified and selected reproductive materials presents some advantages in the context of sustainable forestry. They are collected from populations of known origin, high population size and no phenotypic selected or selected stands. They are also the most frequent of all the existing basic materials (in the EU, for instance, they represent 79.2% of the total basic materials). However, the amount of species under regulation, the presence of different areas for afforestation, and the cost of collecting reproductive material from many different zones, haves shown a reduced availability that amount a major limitation in forestry and ecosystem restoration [33, 34].

We used Peninsular Spain and the Balearic Islands as our model study area, and expanded our study to SPA (basic materials for producing source-identified and selected reproductive materials) in the EU. Our model study area presents a higher tree diversity compared to other areas in Europe [35], as it is a well-known biodiversity hot-spot with a strong interest in species conservation [36]. In this area, there is a long tradition of multi-purpose forest plantations with different goals [37]: restoration, protection and, to a lesser extent, production [38]. There is a preference for source-identified and selected reproductive materials, from a highly diverse species pool [39], that represent 97% of the total reproductive material produced annually in Spain. The other two categories (qualified and tested materials) represent 3% of the basic materials produced for all the species.

We used the defined regions of provenance [40] and deployment zones [41] to evaluate two seed-sourcing strategies: Strict-sense local and Wide-sense local, taking into account the land-scape level, and considering the reported scale of gene flow in forest tree species [42]. To do so, we analysed 40 native forest tree species with actual trading of reproductive materials in Spain, and identified the regions of provenance that were climatically suitable for each species and deployment zone. We also defined the availability of SPA. Our objectives were to determine the richness of the pools for each of the two seed-sourcing strategies, per species and deployment zone, and the relationship among richness, the use of the species reproductive material and the climatic variables of the deployment zones. To conclude, we further discuss the application of local seed-sourcing for afforestation, restoration and reforestation, in a highly diverse environment such as the Mediterranean forests, and its application in other areas in Europe in a context of sustainable forestry.

Material and methods

Species and pools

We considered 40 species (Table 1) regulated for marketing and trading of forest reproductive material in Spain. They were classified depending on whether that material was mainly used for restoration, for protection and to a lesser extent, in productive plantations according to the published national guidelines [38].

Table 1. Species considered in the study.

Code ¹	Species	Use ²	Regpro ³	Code	Species	Use	Regpro
Aal	Abies alba Mill.	RE	6ª	psy	Pinus sylvestris L.	PR,PT	19 ^a
api	Abies pinsapo Boiss.	RE	3ª	pun	Pinus uncinata Ram. ex DC.	PT	6ª
apl	Acer platanoides L.	RE	7 ^d	pav	Prunus avium L.	PR,RE	33 ^d
aps	Acer pseudoplatanus L.	RE	19 ^d	qca	Quercus canariensis Willd.	RE	5 ^a
aun	Arbutus unedo L.	RE	47 ^d	qco	Quercus coccifera L.	RE	33 ^d
bpu	Betula pubescens Ehrh.	RE	23 ^d	qfa	Quercus faginea Lam.	RE	27 ^a
csa	Castanea sativa Mill.	PR	37 ^d	qil	Quercus ilex L.	PR,RE	28 ^a
fsy	Fagus sylvatica L.	PT	18 ^a	qpe	Quercus petraea (Matt.) Liebl.	RE	13 ^a
fex	Fraxinus excelsior L.	RE	17 ^d	qpu	Quercus pubescens Willd.	RE	6 ^a
iaq	Ilex aquifolium L.	RE	30 ^d	qpy	Quercus pyrenaica Willd.	RE	28 ^a
jre	Juglans regia L.	PR	39 ^d	qro	Quercus robur L.	RE	11 ^a
jco	Juniperus communis L.	RE	31 ^d	qsu	Quercus suber L.	PR,RE	25 ^a
jox	Juniperus oxycedrus L.	RE	45 ^d	sar	Sorbus aria (L.) Crantz	RE	31 ^d
jph	Juniperus phoenicea L.	RE	36 ^d	sau	Sorbus aucuparia L.	RE	22 ^d
jth	Juniperus thurifera L.	PT	28 ^d	tga	Tamarix gallica L.	RE	28 ^d
oeu	Olea europea Brot.	RE	45 ^d	tba	Taxus baccata L.	RE	26 ^d
pha	Pinus halepensis Mill.	PT	20 ^a	tco	Tilia cordata Mill.	RE	14 ^d
pni	Pinus nigra Arn.	PR,PT	15 ^a	tpl	Tilia platyphyllos Scop.	RE	18 ^d
ppa	Pinus pinaster Aiton.	PR,PT	29 ^a	ugl	Ulmus glabra Huds.	RE	21 ^d
ppe	Pinus pinea L.	PR,PT	12ª	umi	Ulmus minor Mill. s.l.	RE	46 ^d

¹Code: According to the Commission Regulation (EC) No 1597/2002 of 6 September 2002.

https://doi.org/10.1371/journal.pone.0278866.t001

²Use: from Peman et al. (2013). RE: restauration, PR: production, PT: protection.

³Regpro: number of regions of provenance of the species

⁽aAgglomerative method,

^dDivisive method).

Distribution data were available for each species from the Spanish National Inventory and Spanish Forest Map (1/25000) and transformed into species presence/absence data in a grid of 1x1 km². We considered natural populations–i.e., excluding plantations- in the Spanish Iberian Peninsula and Balearic Islands, excluding plantations, following each species' regions of provenance [40].

Deployment zones were established by a division of Spain into continuous regions with similar environmental characteristics and following a biogeographical classification [43], resulting in fifty units [41, 44] (S1 Fig).

Regions of provenance, are the basic marketing units for source- identified and selected materials according to EU regulations [30]. The regions were defined for all the species considered in this study (Table 1) using two different methodologies: agglomerative and divisive [40]. The agglomerative method (see [45]), groups each species populations with similar ecological, phenotypic or genetic characteristics. These species-specific regions of provenance were defined for 17 main forest species: Abies (2), Fagus (1), Pinus (6) and Quercus (8). The divisive method splits the territory into continuous regions with similar environmental characteristics using a biogeographical classification [41]. This method was applied for the remaining 23 species in the in the study and resulted in regions that coincided with the deployment zones and were not species-specific.

Seed sourcing strategies. A *Strict-sense local seed-sourcing strategy* (SSL) was defined for each deployment zone as the use for each species of autochthonous populations pertaining to the same region of provenance. *Wide-sense local seed-sourcing strategy* (WSL) was defined as the use, in each zone, of regions of provenance with a suitable climatic niche.

Niche modelling was used to obtain the climate-predicted *distribution* of the species based on actual presence and climate (assuming no restrictions to dispersal and no human influence) (see [39] and S1 Annex). Eight climatic variables already used for the niche modelling of different species in Spain [46] were used: Rainfall (P, mm), summer precipitation (SP, mm), winter precipitation (WP), Dry period, considered when P<2T (DP, months), Frost period, considered when T<0°C (FP, months), Annual mean temperature (TM, Celsius degrees), Mean of the Maximum temperatures of the hottest month (MXHM, Celsius degrees) and Mean of the minimum temperatures of the coldest month (*MNCM*, Celsius degrees). Climatic data was obtained for a 1 km² square grid. A similarity index [47] was obtained based on the Mahalanobis climatic distance (using the same set of climatic variables as in the niche modelling estimation, see S1 Annex), comparing the points present both in the region of provenances and each deployment zone. Based on this similarity, the regions of provenance were classified as suitable or not for each deployment zone.

A dataset was created including the regions of provenance identified for each of the 40 species and 50 deployment zones using the SSL and WSL strategies (DOI: 10.5281/zenodo. 7157589).

Seed sourcing pools. We defined eight seed-sourcing pools from this database by combining the two seed-sourcing strategies SSL and WSL at two levels (species and genetic) and considering the availability of SPA in the national register.

The *species pool* was defined as the set of species for each deployment zone following each strategy. Similarly, the *genetic pool* was defined as the set of regions of provenance defined by each strategy.

The available pool was defined as those regions of provenance with at least one SPA in the national register (source-identified or selected categories). The National Register held by the National Authority includes the existing approved basic materials (https://www.miteco.gob.es/es/biodiversidad/temas/recursos-geneticos/geneticos-forestales/rgf_catalogo_materiales_base. aspx; data accessed 31/12/2021).

Data on reproductive material. The actual use of the species in afforestation and reforestation programs was defined by the following variables: Number of region of provenances of the species (*regpro*); Mean number of SPA (entries or accessions in the national register) by region of provenance (*spa*); Afforested area/year in ha (*aff_su*); Ratio of public afforestation to total afforested area (*aff_pu*); Ratio afforested area for protection to total afforested area (*aff_pr*); Source-identified and selected reproductive material by year in number of plants (*frm_si*) and Qualified and tested reproductive material by year in number of plants (*frm_qt*). Data on afforestation and production of reproductive material were obtained as the mean for the period 2005–2016 (last year with available data) using on the Spanish forestry statistics.

We also calculated the percentage of deployment zones with endangered local populations (*recgen*) for each species, following the criteria in the Spanish Strategy for the conservation and sustainable use of forest genetic resources [40, 48]. This value was used as a proxy to the genetic risk of transferring materials to a given zone.

Pool of basic material by country in the EU. We computed the available pool of basic material of the species in our study compiled for each EU country. The FOREMATIS database (https://ec.europa.eu/forematis/) includes the location of the approved basic material, type, origin and purpose by species and country (Community List of Approved Basic Material for the Production of Forest Reproductive Material [48]). We obtained the mean and the harmonic mean of the basic material for all the species in each country (equivalent to an effective number per species) for two classes: SPA (seed sources and stands) and improved basic for qualified and tested categories (seed orchards, clones, parent of families).

Statistical analysis

Richness of seed sourcing pools per deployment zone. We obtained the richness of the eight pools (already defined for each deployment zone) and analyzed the climatic variables related to the richness increment between Strict-sense and Wide-sense local for the different categories (species, genetic and available genetic). A stepwise linear selection model was applied (lm function, stepAIC option in R), to select the variables explaining the relationship among richness increment pools and climatic variables. The procedure start with a saturated model, and the least significant variables are removed until no further decrease in the Bayesian Information Criterion (BIC). Non-significant variables ($\alpha > 0.05$) were also removed from the final selected model. Mean values of the climatic variables used in niche modelling, and the altitude for each deployment zone, were computed based on the 1 km grid, and standardized by the mean and standard deviation across deployment zones.

Relationship among seed sourcing pools and use of the species. We explored the relationship among the set of variables describing the seed sourcing pools by species (richness of the genetic and available genetic pools, number and harmonic mean of deployment zones -equivalent to the effective number of deployment zones- according to the frequency of the species' present in the region), and the variable set describing the actual deployment of reproductive materials (see above). A canonical correspondence analysis (CCA) [49], was performed and the biplot of variables and species was used to explore the relationship among these two sets of variables.

Additionally, we explored the relationship among the variables related to the seed sourcing pools and the variables describing the use of the species for two main type of species depending on the method-agglomerative of divisive- for establishing the regions of provenance. A T-test for two means with unknown population standard deviations (t.test function in the stat package in R), was computed from the means and standard deviation of each of the two groups of species.

Basic materials in the EU. We compared by ANOVA (Im function in R) the values of the number of species under regulation in each country from the 40 considered in the study, and the mean values for the SPA and improved basic materials for different EU regions (Nordic, Western Central, Eastern Central, Western South, Eastern South).

All statistical analyses were made within the R environment (R Core Team, 2015), with the packages corrplot [50], Hmisc [51], CCA [49], and MASS [52].

Results

Seed sourcing pools by deployment zone

The species differed in their range of distribution, and therefore in the local materials available for each deployment zone. The richness of the species pool (for both SSL and WSL strategies) varied greatly among deployment zones. The mean value of the SSL species pool was 20.2 ± 7.6 (range from 4 to 36 species) with a slight increment for the WSL species pool to 26.9 ± 8.1 (range from 8 to 39) (Fig 1 and S1 Table).

When considering the genetic pools, richness increased greatly for both seed-sourcing strategies: to 27.7 ± 12.6 (range from 4 to 63) for SSL and to 114.1 ± 57.8 (range from 10 to 245) for WSL. However, the available genetic pool, (SPAs in the national register), for all species, decreased to 17.7 ± 10.1 (60.2% of the total) for SSL and 57.6 ± 33.9 (48.4% of the total) for WSL strategy.

The increment of richness among SSL and WSL pools showed a climatic trend mainly associated to temperature for the three levels considered: species, genetic and available genetic (Table 2), with a reduction in the areas with higher temperatures, in southern Spain.

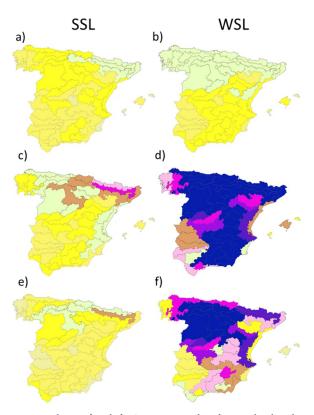


Fig 1. Richness of pools for Strict-sense and Wide-sense local seed-sourcing strategies for the species (a, b), genetic (c, d) and available genetic pools (e, f). BDLJE CC-BY 4.0 ign.es, DR miteco.gob.es.

https://doi.org/10.1371/journal.pone.0278866.g001

Seed sourcing genetic pools per species

The richness per deployment zone of genetic pools differed greatly between SSL and WSL strategies: 1.5 ± 0.7 and 4.0 ± 1.6 , respectively. Moreover, the available pools were significantly lower (1.0 ± 0.8 and 2.1 ± 0.9 respectively). For the 40 species, only 41% of the regions of provenance had at least one basic material (source-identified or selected categories) in the national register, with a mean value of 5.8 basic materials per region of provenance (S2 Table).

Nevertheless, the pools can be used in a significantly higher number of deployment zones (133% as an average increment, from 25.3 ± 12.6 in the Strict-sense local to 33.6 ± 14.1 in the Wide-sense local). This value is reduced when considering availability (15.1 ± 10.0 and 26.2 ± 14.1 respectively).

The canonical correlation analysis showed a trend in the biplot of species on the two first components based on variables related to the use of forest reproductive material, with a transition from those species more extensively used to those used mostly for ecological restoration (Fig 2 and S3 Table). The species with an agglomerative method have a lower number of regions of provenance, a higher amount of SPAs, higher afforested surface, and a higher production of forest reproductive material, including also a higher number of endangered populations (S4 Table).

When comparing the SPA pool richness in Europe, there were large differences among countries (Figs 3 and 4), with a higher number of species and higher number of SPA in southern countries, but the variability among countries within a region was still quite high.

Discussion

We compared the effects of the potential genetic diversity (proxied by the number of suitable and available basic materials) of two alternative local seed sourcing strategies for 40 species in Spain.

Table 2. Regression analysis for the increment of richness for different seed sourcing pools (strict-sense local and Wide-sense local) per deployment zone and climatic variables. Variables retained after a stepwise selection model starting with the eight standardized variables.

					Analysis of variance			
	Estimate	Std. Error	t-value	Pr(> t)	M. Squares	F-value	Pr(>F)	
Species pool								
ALT	-23.642	0.9222	-2.564	0.014025	31.276	7.1488	0.010640	
P	19.680	0.5541	3.551	0.000961	42.224	9.6509	0.003387	
MXHM	-39.996	0.7162	-5.585	1.57e-06	42.865	9.7974	0.003175	
MNCM	-59.040	11.896	-4.963	1.20e-05	107.765	24.6316	1.2e-05	
Genetic pool								
P	-14.999	7.379	-2.033	0.04801	1080	0.9254	0.341203	
SP	-17.473	9.397	-1.859	0.06952	12337	10.5714	0.002179	
TM	-76.989	11.677	-6.593	4.09e-08	47407	40.6211	8.696e-08	
FRO	-29.741	9.189	-3.237	0.00227	12225	10.4754	0.002273	
Available Genetic	pool							
P	-9.846	3.645	-2.702	0.00976	1671.1	6.3874	0.015161	
TM	-67.208	21.459	-3.132	0.00309	17157.3	65.5788	2.918e-10	
FRO	-11.473	4.402	-2.606	0.01245	2188.7	8.3658	0.005923	

ALT: Altitude, P: rainfall, TM: Mean annual Temperature, SP: Summer Precipitation, FRO: Frost period, MXHM: Maximum temperature of the hottest month; MNCM: Minimum temperature of the coldest month. Arid Period was not retained for any case

https://doi.org/10.1371/journal.pone.0278866.t002

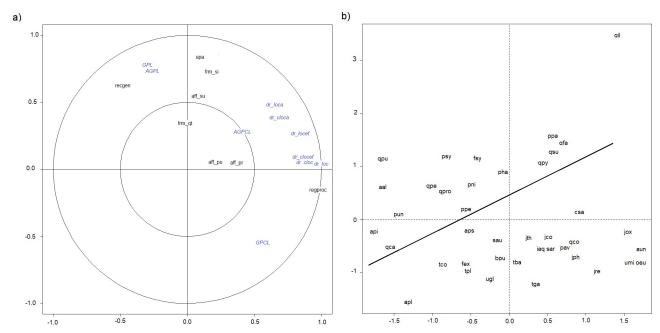


Fig 2. Biplot of the two first components of a CCA analysis on variables related to the seed sourcing pools and variables related to the use of reproductive material for each of the 40 studied species.

https://doi.org/10.1371/journal.pone.0278866.g002

Our analysis is based on deployment zones and regions of provenance in contrast with other studies that work at a finer scale, i.e. the population level (e.g. [26, 53]), because we need to consider the scale at which genetic variation has implications in the use of genetic resources. Regions of provenance are defined using different criteria among and within countries [54], usually within genetic groups defined by neutral markers [55–57]. These regions are in the same order of magnitude as deployment zones. For autochthonous or indigenous populations (if the origin is known), region of provenance can be considered as that of populations within gene-flow distance and therefore suitable for local seed-sourcing. This is in agreement with the results on the response curves of different species which show a range where the populations are close to an optimum [13]. Accordingly, we can assume that the scale at which we define the separate genepools and the deployment zones are adequate for sustainable forestry and restoration activities.

We found that the Strict-sense local species pool was quite high in most of the deployment zones (mean value of 22), with wide genetic pools for those species. We assumed that climate prediction is a good estimate for the performance of the material when we considered Widesense local seed sourcing in our study, that is, that the material from more ecologically similar procurement zones would be preferred for a given deployment zone [see 58,59]. This is a more generalized local seed sourcing method than can be implemented easily for many different species. By using this strategy, the number of basic materials expected to be adapted to the local conditions increased from an average of 1.4 per species and deployment zone to 2.3. This strategy is still rather conservative, discarding recommendations of climatically distant basic materials for a given deployment zone.

We found an expectable climatic trend in the richness increment of the pools, associated with higher annual rainfall (or reduction of drought period) and temperatures. This is a general trend for species diversity, where annual precipitation and mean annual temperature play key favorable roles [60]. The species pool seems large enough for most of the goals of the

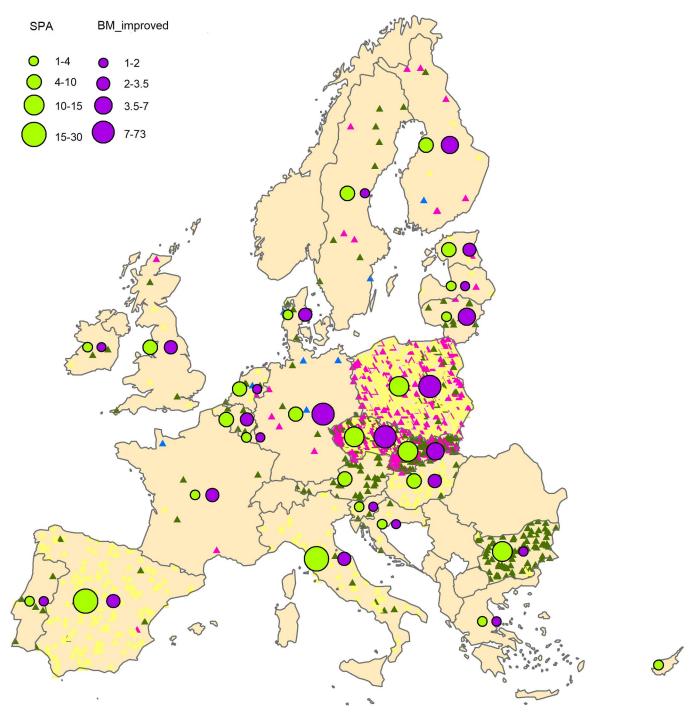


Fig 3. Basic material in the EU for the 40 analyzed species. SPA: harmonic mean of number of seed production areas by country; BM_improved: harmonic mean of number of basic material for qualified and tested categories. Color triangles: Proportional representation of different source categories of forest reproduction material (yellow triangles are identified, green ones are selected, pink ones are qualified and blue ones are tested). Source of basic material: www.forematis.eu. Political map of Europe: https://es.m.wikipedia.org/wiki/Archivo:Political_Map_of_Europe-en.svg. CC BY 4.0.

https://doi.org/10.1371/journal.pone.0278866.g003

plantations (restoration, revegetation, afforestation), allowing the establishment of mixed forests with species differing in functional traits [61] in order to obtain multifunctional forests [62], something particularly relevant in the Mediterranean region.

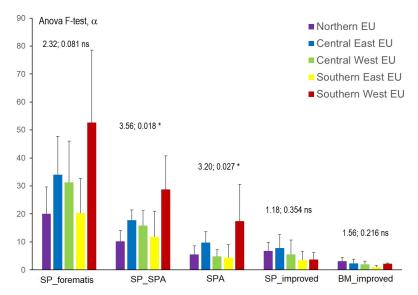


Fig 4. Comparison of basic material in EU regions. SP_Forematis: number of species for each country with approved basic materials; SP_SPA: number of species for each country with basic materials included in this study; SPA: harmonic mean of number of SPA; SP_improved: number of species for each country with approved basic material for the production of qualified and tested reproductive material (excluding clones and parent of families), BM_improved: harmonic mean of number of basic material for the production of qualified and tested reproductive material. Anova F-test and significance level (α) is included for each of the variables analyzed. Source www.forematis.eu. CC-BY 4.0.

https://doi.org/10.1371/journal.pone.0278866.g004

Despite the widespread use of Strict-sense local seed sourcing in many countries and regions, there are relevant concerns about this strategy. Local is not always the best option, for different reasons: the absence of native forest species in the deployment zone is a first obvious one. Moreover, low performance or bad adaptation of the local material has also been reported in some cases [63–66], deriving from reduced population size [67], poor seed crops, low genetic diversity (neutral or adaptive), or unfavorable genetic correlations due to trade-offs among traits in highly contrasting ecosystems [68].

Our study also indicates how relevant the availability–linked to the existence of previously approved SPA–can be a limiting factor decreasing the potential richness of genetic pools. Despite the important effort made in Europe for approving SPA for many important forest species, not all the regions of provenance are listed in the national registers. This drawback makes often ecosystem restoration activities to be often constrained by a lack of the desired or suitable basic material [33], legal restrictions for seed collection in endangered populations and species [69], and biological constraints such as masting, conservation of seeds and fruits and reduced population size.

The implementation of alternative seed sourcing strategies (*eg.* assisted migration, admixture, climate-adjusted) to maximize the future adaptation and resilience of our forests [19, 66, 70, 71] is not feasible in the short term for most of the species under consideration in this study because we lack precise information about the ecological and evolutionary implications derived from the movement of genetic resources in the landscape [72–74]. Furthermore, many local populations will not be in suitable areas in the future, as in the case of Southern populations of Scots pine [66] or in Maritime pine or Aleppo pine [59], and they do not have any source from southern populations with similar expected future conditions. On one hand, an integrative approach aimed at developing suitable restoration materials for a given area [75] based on their out-planting performance requires well designed field experiments with representative planting stock quality [34, 76]. On the other hand, the rapid development of genomic

research can contribute to optimizing seed sourcing strategies by increasing our knowledge of local adaptation and of the main drivers determining the patterns of variation. The existing genetic characterization by molecular markers and/or field trials (comparative common gardens) is clearly insufficient to improve seed sourcing strategies for most of the species. Obtaining this information is costly in money and time (particularly if it involves robust field experimental settings with multiple sites) since fitness assessment in forest tree species requires long-term experimentation [77].

A growing realization about the importance of SPA for producing reproductive material is the general trend in Europe, where it already represents 84% of the basic materials approved. In Nordic and Central East European countries other basic materials (for producing improved materials) have more importance for some of the studied species, and the election of planting stock has to be based on *improved /cultivar seed sourcing strategies* [26, 78–80] not considered in our study. SPA could provide material for restoration and sustainable forestry [81] and restoration [7]. In order to improve the SPA network, it would be necessary to improve the information on the characteristics of the SPA, as the actual information is mostly limited to the location, and to extend to define deployment zones for the different materials based in strategies aimed to sustainable use and restoration to facilitate the election by users.

Conclusions

Local seed sourcing strategies should be the basis for sustainable forestry until more scientifically contrasted information is available. Strict-sense local and Wide-sense local seed sourcing strategies provide sufficient species and provenance pools for deployment zones. These pools show a climatic gradient associated with higher annual rainfall and temperatures. The availability of SPA from specific sources reduces the pools for the different deployment zones, despite the huge efforts done in approving SPAs in Spain in particular and the EU in general. Therefore, it will be necessary to increase planning efforts and offer reproductive materials from specific regions to meet demand.

Supporting information

S1 Annex. Niche modelling and climatic suitability of the species and genetic pools. (DOCX)

S1 Fig. Deployment regions defined in Spain (from García del Barrio et al. 2001. Regiones de identificacion y utilizacion de material forestal de reproduccion. Serie Cartografica. Madrid: MAPA). BDLJE CC-BY 4.0 ign.es, DR miteco.gob.es. (DOCX)

S1 Table. Richness for the seed sourcing pools by deployment zone (DZ). (DOCX)

S2 Table. Richness of seed sourcing pools and deployment zone for 40 forest tree species in Spain.

(DOCX)

S3 Table. Data on the use of forest reproductive material of 40 forest tree species in Spain. (DOCX)

S4 Table. T-test for different variables related to the genetic pools and use of forest reproductive material of the species, comparing species with regions of provenance delineated using the divisive *vs.* agglomerative methods.

(DOCX)

Acknowledgments

We thank F. J. Auñón and D. Sánchez de Ron (ICIFOR-INIA, CSIC) for their contribution in niche modelling of the different species included in this study. P.C. Grant revised the English version.

Author Contributions

Conceptualization: Ricardo Alía, Eduardo Notivol, José Climent, Juan Majada.

Data curation: Diana Barba.

Formal analysis: Ricardo Alía, Eduardo Notivol, José Manuel García del Barrio.

Funding acquisition: Ricardo Alía, José Climent, Felipe Pérez.

Validation: Felipe Pérez.

Writing – original draft: Ricardo Alía.

Writing – review & editing: Eduardo Notivol, José Climent, Felipe Pérez, Diana Barba, Juan Majada, José Manuel García del Barrio.

References

- Campos P, Caparrós A, Oviedo JL, Ovando P, Álvarez-Farizo B, Díaz-balteiro L, et al. Bridging the gap between national and ecosystem accounting application in andalusianforests, Spain. Ecol Econ. 2019; 157: 218–236. https://doi.org/10.1016/j.ecolecon.2018.11.017
- 2. European_Comssion. New EU Forest Strategy for 2030. COM(2021) 572 Final. 2021; 27.
- Konnert M, Fady B, Gömöry D, A 'hara S, Wolter F, Ducci F, et al. Use and transfer of forest reproductive material in Europe in the context of climate change. Rome, Italy: Bioversity International; 2015.
 Available: http://www.euforgen.org/fileadmin/templates/euforgen.org/upload/Publications/Thematic_publications/EUFORGEN_FRM_use_transfer_web.pdf
- Ruiz-Jaen MC, Mitchell Aide T, Aide TM. Restoration success: How is it being measured? Restor Ecol. 2005; 13: 569–577.
- 5. Stanturf JA. What is forest restoration? In: Stanturf J.A., Madsen P, editor. Restoration of Boreal and Temperate Forests. Boca Raton: CRC Press; 2005. pp. 3–11.
- Kettenring KM, Mercer KL, Reinhardt Adams C, Hines J. Application of genetic diversity-ecosystem function research to ecological restoration. J Appl Ecol. 2014; 51: 339–348. https://doi.org/10.1111/1365-2664.12202
- Zinnen J, Broadhurst LM, Gibson-Roy P, Jones TA, Matthews JW. Seed production areas are crucial to conservation outcomes: benefits and risks of an emerging restoration tool. Biodivers Conserv. 2021; 30: 1233–1256. https://doi.org/10.1007/s10531-021-02149-z
- 8. van Buijtenen JP. Fundamental Genetic Principles. In: Fins AT, Brotschol J V, Friedman L, editors. Handbook of Quantitative Forest Genetics. Dordrecht: Kluwer Academic Publishers: 1992. pp. 29–68.
- Koskela J, Vinceti B, Dvorak W, Bush D, Dawson IK, Loo J, et al. Utilization and transfer of forest genetic resources: a global review. For Ecol Manage. 2014; 333: 22–34. https://doi.org/10.1016/j.foreco.2014.07.017
- Graudal L, Aravanopoulos FA, Bennadji Z, Changtragoon S, Fady B, Kjær ED, et al. Global to local genetic diversity indicators of evolutionary potential in tree species within and outside forests. For Ecol Manage. 2014; 333: 35–51. https://doi.org/10.1016/j.foreco.2014.05.002
- Fady B, Cottrell J, Ackzell L, Alía R, Muys B, Prada A, et al. Forests and global change: what can genetics contribute to the major forest management and policy challenges of the twenty-first century? Reg Environ Chang. 2016; 16: 927–939. https://doi.org/10.1007/s10113-015-0843-9
- Kawecki TJ, Ebert D. Conceptual issues in local adaptation. Ecol Lett. 2004; 7: 1225–1241. https://doi.org/10.1111/j.1461-0248.2004.00684.x
- Savolainen O, Pyhajarvi T, Knurr T, Pyhäjärvi T, Knürr T. Gene flow and local adaptation in tees. Annu Rev Ecol Evol Syst. 2007; 38: 595–619. https://doi.org/10.1146/annurev.ecolsys.38.091206.095646

- Alberto FJ, Aitken SN, Alía R, González-Martínez SC, Hänninen H, Kremer A, et al. Potential for evolutionary responses to climate change—evidence from tree populations. Glob Chang Biol. 2013; 19: 1645–1661. https://doi.org/10.1111/gcb.12181 PMID: 23505261
- 15. Mátyás C. Climatic adaptation of trees: Rediscovering provenance tests. Euphytica. 1996; 92: 45–54.
- Breed MF, Stead MG, Ottewell KM, Gardner MG, Lowe AJ. Which provenance and where? Seed sourcing strategies for revegetation in a changing environment. Conserv Genet. 2013; 14: 1–10. https://doi.org/10.1007/s10592-012-0425-z
- 17. IUCN. Afforestation and Reforestation for Climate Change Mitigation: Potentials for Pan-European Action. Quality. Warsaw: IUCN Programme Office for Central Europe; 2004.
- Havens K, Vitt P, Still S, Kramer AT, Fant JB, Schatz K. Seed Sourcing for Restoration in an Era of Climate Change. Nat Areas J. 2015; 35: 122–133. https://doi.org/10.3375/043.035.0116
- Prober SM, Byrne M, Mclean EH, Steane DA, Potts BM, Vaillancourt RE. Climate-adjusted provenancing: a strategy for climate-resilient ecological restoration. Front Ecol Environ. 2015; 3: 65. https://doi.org/10.3389/fevo.2015.00065
- Parker WH. Focal point seed zones: site-specific seed zone delineation using geographic information systems. Can J For Res. 1992; 22: 267–271.
- Wang T, Hamann A, Yanchuk A, O'Neill G a., Aitken SN. Use of response functions in selecting lodge-pole pine populations for future climates. Glob Chang Biol. 2006; 12: 2404–2416. https://doi.org/10. 1111/j.1365-2486.2006.01271.x
- 22. Lindgren D, Ying CC. A model integrating seed source adaptation and seed use. New For. 2000; 20: 87–104.
- 23. Wang T, O'Neill G a, Aitken SN. Integrating environmental and genetic effects to predict responses of tree populations to climate. Ecol Appl. 2010; 20: 153–63. Available: https://doi.org/10.1890/08-2257.1
- 24. Farjat A, Reich BJ, Guinness J, Whetten R, Mckeand S, Isik F. Optimal Seed Deployment under Climate Change using Spatial Models: Application to Loblolly Pine in the Southeastern US. J Am Stat Assoc. 2017; 112: 909–920. https://doi.org/10.1080/01621459.2017.1292179
- 25. Ramalho CE, Byrne M, Yates CJ. A Climate-Oriented Approach to Support Decision-Making for Seed Provenance in Ecological Restoration. Front Ecol Evol. 2017; 5: 1–10. https://doi.org/10.3389/fevo. 2017.00095
- Berlin M. Planter's guide—a joint decision support tool for Scots pine plant material in Sweden and Finland. 2019. Available: https://www.skogforsk.se/english/news/2019/common-scots-pine-deploymentrecommendations-for-sweden-and-finland/
- Hamann A, Koshy MP, Namkoong G, Ying CC. Genotype-environment interactions in Alnus rubra: developing seed zones and seed-transfer guidelines with spatial statistics and GIS. For Ecol Manage. 2000; 136: 107–119.
- Zobel M, Maarel E van der, Cecilia D. Species pool: the concept, its determination and significance for community restoration. Appl Veg Sci. 1998; 1: 56–66.
- Jones TA, Monaco TA. A role for assisted evolution in designing native plant materials for domesticated landscapes. Front Ecol Environ. 2009; 7: 541–547. https://doi.org/10.1890/080028
- Council EU. Council Directive 1999/105/EC of 22 December 1999 on the marketing of forest reproductive material. Off J Eur Communities. 2000: 11: 17–40.
- Nanson A. The New OECD Scheme for the Certification of Forest Reproductive Materials. Silvae Genet. 2001; 50: 181–187.
- **32.** Mangold RD, Bonner FT. Certification of Tree Seeds and Other Woody Plant Materials Certification of Tree Seeds and Other Woody Plant Materials. In: Bonner FT, Karrfalt RP, editors. The Woody Plant Seed Manual. USDA.; 2008. pp. 117–124.
- **33.** Jalonen R, Valette M, Boshier D, Duminil J, Thomas E. Forest and landscape restoration severely constrained by a lack of attention to the quantity and quality of tree seed: Insights from a global survey. Conservation Letters Wiley-Blackwell; Jul 1, 2018 pp. 1–9. https://doi.org/10.1111/conl.12424
- McCormick ML, Carr AN, Massatti R, Winkler DE, De Angelis P, Olwell P. How to increase the supply of native seed to improve restoration success: the US native seed development process. Restor Ecol. 2021; 29: 1–9. https://doi.org/10.1111/rec.13499
- Scarascia-Mugnozza G, Oswald H, Piussi P, Radoglou K. Forests of the Mediterranean region: Gaps in knowledge and research needs. For Ecol Manage. 2000; 132: 97–109. https://doi.org/10.1016/S0378-1127(00)00383-2
- Myers N, Mittlermeier RA, Mittlermeier CG, da Fonseca GA, Kent J. Biodiversity hotspots for conservation priorities. Nature. 2000; 403: 853–858. https://doi.org/10.1038/35002501 PMID: 10706275

- Fernández-Ondoño E, Serrano LR, Jiménez MN, Navarro FB, Díez M, Martín F, et al. Afforestation improves soil fertility in south-eastern Spain. Eur J For Res. 2010; 129: 707–717. https://doi.org/10.1007/s10342-010-0376-1
- Peman J, Navarro-Cerrillo RM, Nicolás JL, Prada MA, Serrada R. Produccion y manejo de semillas y plantas forestales. Vol II. Madrid: OAPN, MAGRAMA; 2013.
- García del Barrio JM, Auñón F, Sánchez de Ron D, Alía R. Assessing regional species pools for restoration programs in Spain. New For. 2013; 44: 559–576. https://doi.org/10.1007/s11056-013-9363-y
- 40. Alía R, Garcia del Barrio JM, Iglesias S, Mancha JA, de Miguel J, Nicolás JL, et al. Regiones de procedencia de especies forestales de España. Madrid, Spain: O.A. Parques Nacionales; 2009.
- 41. García del Barrio JM, de Miguel J, Alía R, Iglesias S. Regiones de identificacion y utilizacion de material forestal de reproduccion. Serie Cartografica. Madrid: MAPA; 2001.
- Kremer A, Ronce O, Robledo-Arnuncio JJ, Guillaume F, Bohrer G, Nathan R, et al. Long-distance gene flow and adaptation of forest trees to rapid climate change. Ecol Lett. 2012; 15: no-no. https://doi.org/ 10.1111/j.1461-0248.2012.01746.x PMID: 22372546
- 43. Bunce RGH, Carey PD, Elena-Rossello R, Orr J, Watkins J, Fuller R. A comparison of different biogeographical classifications of Europe, Great Britain and Spain. J Environ Manage. 2002; 65: 121–134. https://doi.org/10.1006/jema.2002.0533 PMID: 12197075
- 44. García del Barrio JM, Iglesias S, Alía R. Regiones de Identificación y Utilización de Material Forestal de Reproducción. Regiones de procedencia en España realizadas por el métodod divisivo. Adenda. Madrid: Ministerio de Medio Ambiente; 2005.
- 45. Catalan G, Gil P, Galera R, Martín S, Agúndez D, Alía R. Las regiones de procedencia de Pinus sylvestris L. y Pinus nigra Arn subsp. salzmanii (Dunal) Franco en España. Madrid: ICONA; 1991.
- 46. Benito-Garzón M, Sánchez de Dios R, Sáinz Ollero H. Predictive modelling of tree species distributions on the Iberian Peninsula during the Last Glacial Maximum and Mid-Holocene. Ecography (Cop). 2007; 30: 120–134. https://doi.org/10.1111/j.2006.0906-7590.04813.x
- Doherty KD, Butterfield BJ, Wood TE. Matching seed to site by climate similarity: Techniques to prioritize plant materials development and use in restoration. Ecol Appl. 2017; 27: 1010–1023. https://doi.org/10.1002/eap.1505 PMID: 28112847
- **48.** Jiménez P, Diaz-Fernandez PM, Iglesias S, Prada MA, Garcia del Barrio JM, Alba N, et al. Strategy for the Conservation and Sustainable Use of Spanish Forest Genetic Resources. Inv Agrar Sist y Rec. 2009: 18: 13–19.
- Gonzalez I, Dejean S, Martin PGP, Baccini A. Journal of statistical software. J Stat Softw. 2008; 23: 1– 14. https://doi.org/10.1002/wics.10
- Wei T, Simko V. R package "corrplot": Visualization of a Correlation Matrix. 2021. Available: https://github.com/taiyun/corrplot
- Harrell FE. Hmisc: A package of miscellaneous R functions. 2021. Available: https://cran.r-project.org/ package=Hmisc
- Ripley B. Support Functions and Datasets for Venables and Ripley's MASS. 2022. Available: http://www.stats.ox.ac.uk/pub/MASS4/
- Lesser MR, Parker WH. Comparison of canonical correlation and regression based focal point seed zones of white spruce. Can J For Res. 2006; 36: 1572–1586.
- 54. Auñon FJ, Garcia del Barrio JM, Mancha JA, de Vries SMG, Alía R. Regions of provenance of European beech (Fagus sylvatica L.) in Europe. In: von Wuehlisch G, Alia R, editors. Genetic resources of European beech (Fagus sylvatica L) for sustainable Forestry. Madrid (España): INIA; 2010. pp. 141–148.
- 55. Rodríguez-Quilón I, Santos-del-Blanco L, Grivet D, Jaramillo-Correa JP, Majada J, Vendramin GG, et al. Local effects drive heterozygosity–fitness correlations in an outcrossing long-lived tree. Proc R Soc B Biol Sci. 2015; 282: 20152230. https://doi.org/10.1098/rspb.2015.2230 PMID: 26631567
- De Kort H, Mergeay J, Vander Mijnsbrugge K, Decocq G, Maccherini S, Kehlet Bruun HH, et al. An evaluation of seed zone delineation using phenotypic and population genomic data on black alder Alnus glutinosa. J Appl Ecol. 2014; 51: 1218–1227. https://doi.org/10.1111/1365-2664.12305
- 57. Bucci G, Vendramin GG. Delineation of genetic zones in the European Norway spruce natural range: preliminary evidence. Mol Ecol. 2000; 9: 923–34. Available: http://www.ncbi.nlm.nih.gov/pubmed/10886655 https://doi.org/10.1046/j.1365-294x.2000.00946.x
- Serra-Varela MJ, Grivet D, Vincenot L, Broennimann O, Gonzalo-Jiménez J, Zimmermann NE. Does phylogeographical structure relate to climatic niche divergence? A test using maritime pine (P inus pinaster Ait.). Glob Ecol Biogeogr. 2015; 24: 1302–1313. https://doi.org/10.1111/geb.12369

- Serra-Varela MJ, Alía R, Daniels RR, Zimmermann NE, Gonzalo-Jiménez J, Grivet D. Assessing vulnerability of two Mediterranean conifers to support genetic conservation management in the face of climate change. Divers Distrib. 2017; 23: 507–516. https://doi.org/10.1111/ddi.12544
- 60. Baudena M, Sánchez A, Georg CP, Ruiz-Benito P, Rodríguez MA, Zavala MA, et al. Revealing patterns of local species richness along environmental gradients with a novel network tool. Sci Rep. 2015; 5: 1–15. https://doi.org/10.1038/srep11561 PMID: 26109495
- 61. Baeten L, Verheyen K, Wirth C, Bruelheide H, Bussotti F, Finér L, et al. A novel comparative research platform designed to determine the functional significance of tree species diversity in European forests. Perspect Plant Ecol Evol Syst. 2013; 15: 281–291. https://doi.org/10.1016/j.ppees.2013.07.002
- 62. Manning P, Van Der Plas F, Soliveres S, Allan E, Maestre FT, Mace G, et al. Redefining ecosystem multifunctionality. Nat Ecol Evol. 2018; 2: 427–436. https://doi.org/10.1038/s41559-017-0461-7 PMID: 29453352
- **63.** Namkoong G. Nonoptimality of local races. Proceedings of the tenth southern Conference on Forest Tree Improvement. Houston, Texas; 1969.
- 64. Leimu R, Fischer M. A Meta-Analysis of Local Adaptation in Plants. PLoS One. 2008; 3: 1–8. https://doi. org/10.1371/journal.pone.0004010 PMID: 19104660
- Jones TA. When local isn't best. Evol Appl. 2013; 6: 1109–1118. https://doi.org/10.1111/eva.12090
 PMID: 24187591
- 66. Notivol E, Santos-del-Blanco L, Chambel R, Climent J, Alía R. Seed sourcing strategies considering climate change forecasts: A practical test in scots pine. Forests. 2020; 11: 1–17. https://doi.org/10.3390/f11111222
- **67.** Robledo-Arnuncio JJ, Alía R, Gil LA. Increased selfing and correlated paternity in a small population of a predominantly outcrossing conifer, Pinus sylvestris. Mol Ecol. 2004; 13: 2567–77. https://doi.org/10.1111/j.1365-294X.2004.02251.x PMID: 15315671
- Ramírez-Valiente JA, Cavender-Bares J. Evolutionary trade-offs between drought resistance mechanisms across a precipitation gradient in a seasonally dry tropical oak (Quercus oleoides). Tree Physiol. 2017; 37: 889–901. https://doi.org/10.1093/treephys/tpx040 PMID: 28419347
- 69. Sánchez-Robles JM, Balao F, García-Castaño JL, Terrab A, Navarro-Sampedro L, Talavera S. Nuclear Microsatellite Primers for the Endangered Relict Fir, Abies pinsapo (Pinaceae) and Cross-Amplification in Related Mediterranean Species. Int J Mol Sci. 2012; 13: 14243–50. https://doi.org/10.3390/iims131114243 PMID: 23203061
- Crowe KA, Parker WH. Using portfolio theory to guide reforestation and restoration under climate change scenarios. Clim Change. 2008; 89: 355–370. https://doi.org/10.1007/s10584-007-9373-x
- Rehfeldt GE, Ying CC, Spittlehouse DL, Hamilton DA. Genetic responses to climate in Pinus contorta: Niche breadth, climate change, and reforestation. Ecol Monogr. 1999; 69: 375–407.
- 72. Ahteensuu M, Lehvävirta S. Assisted Migration, Risks and Scientific Uncertainty, and Ethics: A Comment on Albrecht et al.'s Review Paper. J Agric Environ Ethics. 2014 [cited 7 Feb 2014]. https://doi.org/10.1007/s10806-014-9493-z
- McLachlan JS, Hellmann JJ, Schwartz MW. A framework for debate of assisted migration in an era of climate change. Conserv Biol. 2007; 21: 297–302. https://doi.org/10.1111/j.1523-1739.2007.00676.x PMID: 17391179
- Hamann A, Roberts DR, Barber QE, Carroll C, Nielsen SE. Velocity of climate change algorithms for guiding conservation and management. Glob Chang Biol. 2015; 21: 997–1004. https://doi.org/10.1111/ gcb.12736 PMID: 25310933
- 75. Landis TD, Dumroese RK. Applying the target plant concept to nursery stock quality. Plant Qual—A key to success For Establ Proc CONFORD Conf. 2006; 1–10. Available: http://217.205.94.38/pdf/CofordPlantQuality.pdf/\$file/CofordPlantQuality.pdf#page=7
- Pinto JR, Dumroese RK, Davis AS, Landis TD. Conducting seedling stocktype trials: A new approach to an old question. J For. 2011; 109: 293–299. https://doi.org/10.1093/jof/109.5.293
- Ramírez-Valiente JA, Santos del Blanco L, Alía R, Robledo-Arnuncio JJ, Climent J. Adaptation of Mediterranean forest species to climate: Lessons from common garden experiments. J Ecol. 2021; 1–21. https://doi.org/10.1111/1365-2745.13730
- 78. Chakraborty D, Wang T, Andre K, Konnert M, Lexer MJ, Matulla C, et al. Adapting Douglas-fir forestry in Central Europe: evaluation, application, and uncertainty analysis of a genetically based model. Eur J For Res. 2016; 135: 919–936. https://doi.org/10.1007/s10342-016-0984-5
- Freitas ECS de, Paiva HN de, Neves JCL, Marcatti GE, Leite HG. Modeling of eucalyptus productivity with artificial neural networks. Ind Crops Prod. 2020;146. https://doi.org/10.1016/j.indcrop.2020.112149

- 80. Marcatti GE, Resende RT, Resende MD V., Ribeiro CAAS, dos Santos AR, da Cruz JP, et al. GIS-based approach applied to optimizing recommendations of Eucalyptus genotypes. For Ecol Manage. 2017; 392: 144–153. https://doi.org/10.1016/J.FORECO.2017.03.006
- **81.** Alía R, Alba N, Chambel MR, Barba D, Iglesias S. Genetic quality of forest reproductive materials in Land restoration programmes. In: Vallejo VR, Bautista S, Aaronson J, editors. Innovative approaches in forest restoration. Valencia: CEAM; 2009. pp. 89–103.