

1 Leveraging multispectral and LiDAR UAV to predict
2 individual tree health: a case study of *Viscum album*
3 in Scots pine forests

4 Ortiz-Ayuso, J. ^{1*}, Sancho-Knapik, D. ^{2,3}, Saz, M.A. ^{4,5}, Hoffrén, R. ^{6,7}, Domingo, D. ^{1,7,8}

5

6 ¹ EifAB-iuFOR, Campus de Soria, University of Valladolid, E-42004. Soria, Spain,
7 jorge.ortiz@uva.es, dario.domingo@uva.es

8 ² Agrifood Research and Technology Centre of Aragón (CITA). Avda Montañana 930, 50059
9 Zaragoza, Spain, dsancho@cita-aragon.es

10 ³ AgriFood Institute of Aragon (IA2) – IA2 (CITA-University of Zaragoza). Zaragoza, Spain

11 ⁴ Department of Geography and Regional Planning. Environmental Sciences Institute (IUCA).
12 University of Zaragoza. Pedro Cerbuna 12, 50009 Zaragoza, Spain, masaz@unizar.es

13 ⁵ Environmental Sciences Institute (IUCA). Climate, water, global change, and natural systems
14 group. University of Zaragoza. Zaragoza, Spain.

15 ⁶ Pyrenean Institute of Ecology, Spanish National Research Council, Avda. Montañana 1005,
16 50059 Zaragoza, Spain, rhoffren@ipe.csic.es

17 ⁷ Environmental Sciences Institute (IUCA). GEOFOREST. University of Zaragoza. Zaragoza,
18 Spain.

19 ⁸ Department of Geography. University of Valladolid. Valladolid, Spain.

20

21

22

23

24

25

26 **Abstract**

27 The presence of mistletoe in pine stands has expanded in recent decades, currently
28 threatening Mediterranean forests. Mistletoe outbreaks can make the host trees more
29 vulnerable to intense droughts, which are expected to increase due to climate change. We
30 use multispectral (MS) and LiDAR UAV-derived data to determine *Viscum album* ssp.
31 *austriacum* infestation levels at individual tree level in Scots pine (*Pinus sylvestris* L.)
32 forests. First, spectral and structural differences between four infestation levels were
33 assessed employing Kruskal-Wallis test and post hoc Dunn's test for individual tree
34 crowns. Second, machine learning classification algorithms were applied to evaluate
35 infestation levels at the individual tree scale by comparing or combining UAV-derived
36 datasets. Outcomes revealed significant differences between infestation levels in canopy
37 cover and height based on LiDAR derived metrics. Significant changes in vegetation
38 vigor were also found through spectral and textural metrics. Using two vegetation indices
39 (CIRE and NDVI) an overall accuracy of 0.83 was achieved by applying SVM, while
40 combining a spectral metric (NDRE) and a LiDAR metric (D0) resulted in 0.82 accuracy
41 with SVM. Using only LiDAR variables, we obtained an accuracy of 0.64 with SVM and
42 RF. This approach demonstrates their value for detecting and characterizing
43 morphological changes in up to four levels of mistletoe infestation at individual trees in
44 Mediterranean Scots pine forests, lending support to forest management monitoring.

45 **Keywords:** UAV; multispectral; LiDAR; machine learning; forest health monitoring;
46 mistletoe.

47

48

49

50

51

52

53

54

55 1. Introduction

56 Hemiparasitic plants are considered a biotic factor that affects forest ecosystems
57 worldwide, leading to the decline of forest stands (Sangüesa-Barreda et al., 2013;
58 Dobbertin, 2005). In combination with abiotic factors such as prolonged droughts, prone
59 to increase its frequency under current global warming, they are considered as a
60 contributing stress factor for pine vulnerability (Dobbertin & Rigling, 2006; Sangüesa-
61 Barreda et al., 2013; Allen et al., 2010; Sabrina et al., 2020). In fact, ongoing temperature
62 rise and shifts in precipitation patterns make predicting pest dynamics increasingly
63 challenging (Torres et al., 2021). The monitoring and management of hemiparasitic plants
64 is of great interest in southern distribution limits (Zuber, 2004).

65 *Viscum album* L. is an evergreen hemiparasitic plant with persistent haustorium that is
66 connected to the host tree through one-way water flow, enabling the transfer of
67 photosynthates and nutrients between the host and the parasite (Glatzel & Geils, 2009;
68 Hernández-Jiménez, 2020). The ssp. *austriacum* is the one found on pines such as *P.*
69 *sylvestris*, *P. halepensis*, *P. nigra*, and less frequently on *P. pinaster* and *P. uncinata*
70 (Hernández-Alonso et al., 2001; Zuber, 2004). The presence of mistletoe is common and
71 necessary, as its fruits provide food for birds and its leaves serve as sustenance for insects
72 (Mathiasen et al., 2008). The distribution area of *V. album* ranges from 10° W to 80° E,
73 and from 60° N to 35° S, with the Mediterranean Sea marking the southern boundary and
74 the Atlantic Ocean the western boundary (Zuber, 2004). It appears in areas below 1000
75 meters in altitude, but when exposed to sunlight, it can be found at higher elevations
76 (Zuber, 2004). The presence of mistletoe decreases the vigor and growth of the host and
77 may also exacerbate its water stress during periods of drought (Sangüesa-Barreda et al.,
78 2018). Under water deficit situations, if prolonged over time, can lead to the depletion of
79 the tree's resources and begin to show symptoms of decline or complete death (Sancho-
80 Knapik et al., 2017) as seen in Mediterranean Scots pine forests within the Iberian System.

81 The abundance of mistletoe in tree canopies is generally assessed through *in situ* studies,
82 which involve high human costs. Though, determining the degree of infestation by
83 fieldwork is challenging because mistletoe typically grows in the upper canopy. The
84 integration of remote sensing data, captured at zenith position, offers quantitative insights
85 to complement *in situ* surveys enabling spatially detailed monitoring. Remote sensing is
86 a highly promising data source for monitoring forest health and is continuously evolving
87 (Tymińska-Czabańska et al., 2024; Senf et al., 2018). This method can offer an automated

88 and customized solution for accurately detecting and classifying mistletoe (Sabrina et al.,
89 2020). Mistletoe research has been conducted using data collected from multispectral
90 imagery from satellite as done by Thapa (2013), unmanned aerial vehicle (UAV) at ~30
91 m with hyperspectral sensor, as used by Ančić et al. (2014), and hyperspectral bands and
92 LiDAR data from airborne flights as applied by Barbosa et al. (2016). The combination
93 of high spatial and temporal resolutions, adaptability, and reduced operational expenses
94 makes UAVs a viable substitute to traditional remote sensing platforms (Guimarães et al.,
95 2020).

96 UAVs are rapidly advancing as an innovative technology for monitoring forest ecosystem
97 health (Torres et al., 2021). Ecke et al. 2022 review their application for this purpose,
98 while Missarov et al. (2024) focus on the use of both UAVs and manned aircraft in
99 mistletoe research. Tymińska-Czabańska et al. (2024) aim to assess the probability of
100 mistletoe presence in *Pinus sylvestris* L. stands using UAV and ALS. Maes et al. (2018)
101 analyze interactions between host and mistletoe by UAV-based infrared thermography.
102 Miraki et al. (2021) used RGB data from UAVs (collected during winter and summer
103 flights in mixed broadleaved forests) to derive a CHM, subsequently used to segment
104 individual trees and classify them as infested and non-infested using RGB bands and
105 Random Forest. León-Bañuelos et al. (2020) use RGB UAV-derived data to identify
106 phenological stages of *Arceuthobium globosum* using colorimetric ranges at pixel level
107 (CRPL) algorithm. Miszczyszyn & Wezyk (2022) investigate the suitability of high-
108 resolution RGB and multispectral data from UAVs, along with derived vegetation
109 indices, for monitoring mistletoe in pine stands. They outline a method utilizing machine
110 learning algorithms like SVM and RF, which could significantly advance mistletoe
111 research. Therefore, multispectral data have been used to mistletoe identification. Mejia-
112 Zuluaga et al. (2022) present a Genetic Programming (GP) method for the automated
113 design of a model utilizing multispectral UAV images to identify mistletoe. On the other
114 hand, Missarov et al. (2022) propose a drone LiDAR for mistletoe recognition and
115 monitoring, a technology also employed by Barbosa et al. (2016). These authors used
116 LiDAR to determine the average height of infested trees, the relative height of mistletoe
117 in the tree canopies, combined with fieldwork, and to classify the landscape structure,
118 which helped identify areas where mistletoe is most prevalent. Thapa (2013) used LiDAR
119 to derive a CHM, which was employed to determine the center of the plot and map the
120 individual tree locations within it.

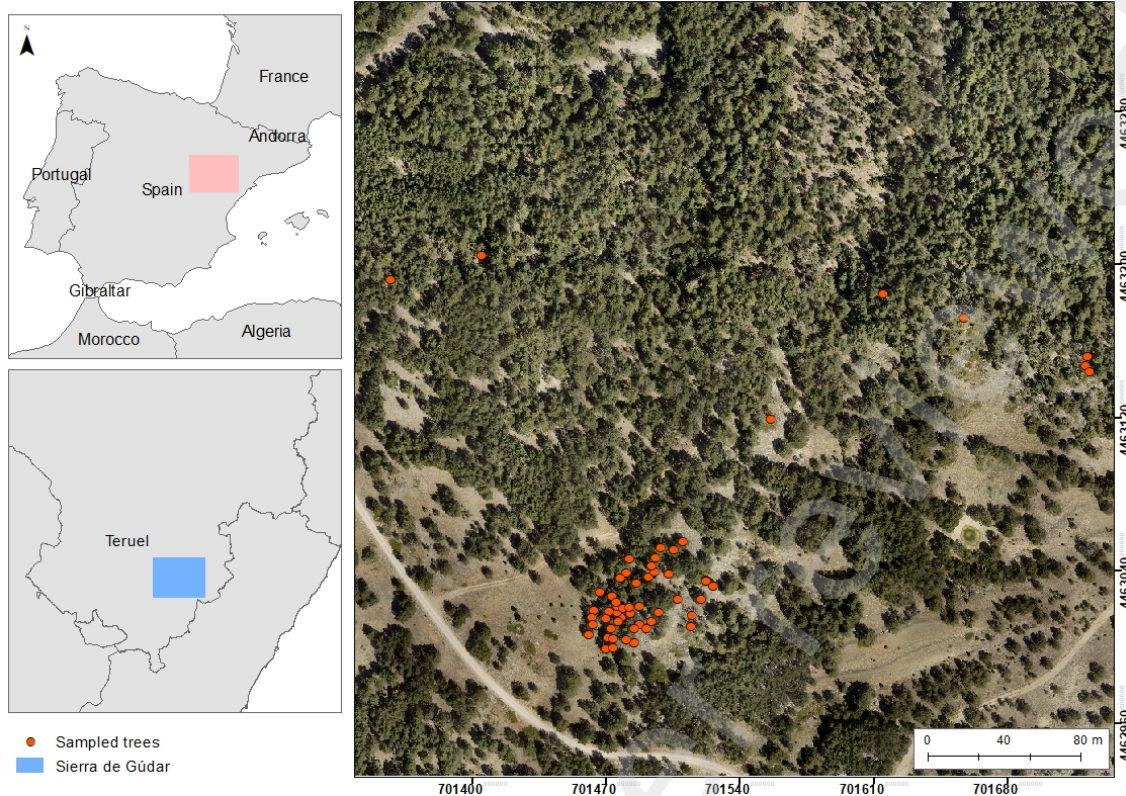
121 In this study, we determine the *Viscum album* ssp. *austriacum* infestation levels at
122 individual tree scale on Scots pine forests by using multispectral and LiDAR UAV-
123 derived data. Our aims are (i) to identify the existence of spectral and structural
124 differences between up to four mistletoe infestation levels and (ii) evaluate the potential
125 of multispectral and LiDAR UAV-derived metrics to classify infestation levels,
126 comparing and combining LiDAR and multispectral UAV-derived datasets.

127 **2. Material and methods**

128 *2.1 Study area and field data collection*

129 The study area is located in Sierra de Gúdar, within the Iberian Range in Teruel region
130 province (Aragón, Spain). The area covers 20 hectares, corresponding to UAV flights
131 equipped with RGB, multispectral and LiDAR sensors within a public-use forest stand.
132 The stand is dominated by Scots pine (*Pinus sylvestris* L.), which constitutes the southern
133 limit of the species' distribution in the western mediterranean and accompanied by
134 *Juniperus* ssp. The area ranges in elevation from 1600 to 1800 m a.s.l., and the mean
135 annual temperature is 9.4°C and annual rainfall averages 700 mm. The main lithology is
136 composed of Early Cretaceous marls and marls limestones.

137 Field data collection was carried out in 55 selected trees (Figure 1). For each tree, a
138 mistletoe infection level (from 1 to 4) was assigned in the field based on expert
139 knowledge. The established levels were determined using a three pair wise including
140 researchers and forest health experts from the Aragon Forest Service.



141

142 **Figure 1.** Location of the research area and spatial distribution of the 55 sampled trees within the
 143 30T HUSE grid of the UTM coordinate system. The orthomosaic was obtained from imagery
 144 captured during a DJI Matrice 300 RTK drone flight using an RGB sensor conducted in April
 145 2023.

146

147 The established levels of mistletoe infection in Scots pine were as follows: Level 1, no
 148 presence of mistletoe or only one clump. Level 2, more than one mistletoe clump but
 149 green needle foliage is more abundant than mistletoe. Level 3, green needle foliage is less
 150 abundant than mistletoe. Level 4, a tree with abundant mistletoe without green needle
 151 foliage or one that is dead. These levels were determined according to infestation criteria
 152 established by the Aragón Region Forest Health Monitoring Network (Gobierno de
 153 Aragón et al., 2020). The first field campaign was conducted in July 2023, involving a
 154 minor sampling based on UAV, followed by a second campaign in May 2024 to re-
 155 evaluate the infestation levels. The trees analyzed were selected based on the existence
 156 of no changes in mistletoe infestation levels and defoliation between UAV data
 157 acquisition in 2023 respect to the re-evaluation field campaign in 2024.

158 *2.2 UAV multispectral and LiDAR data capture and processing*

159 Data collection involved two flight missions using a DJI Matrice 300 platform. The drone
160 platform was fitted with Global Navigation Satellite System (GNSS) and real-time
161 Kinematic (RTK) weighting 6.3 kg. The flights were conducted in April 11, 2023.

162 The first flight utilized the Micasense Altum P camera to acquire multispectral imagery
163 across RGB and near infrared wavelengths. The camera weighs 577 g and has dimensions
164 of 11.0 x 8.0 x 6.9 cm. The flight altitude was set at 70 m, with a 20 m swath width, and
165 a speed of 5 m/s was set. The ground sample distance (GSD) was 3 cm, and a field of
166 view of 50° HFOV x 38° VFOV. The spectral bands were: blue (475 nm), green (560
167 nm), red (668 nm), red edge (717 nm), and NIR (842 nm).

168 The second flight equipped a Phoenix Aerial LiDAR Scout Ultra system with Velodyne
169 sensors to generate a LiDAR point cloud, in addition to capture an RGB image. LiDAR
170 sensor weighs 2.2 kg, and has dimensions of 18.5 x 11.6 x 11.6 cm. It includes an inertial
171 measurement unit for accurate management of flight settings. The scan rate was 600 k
172 points/s, with up to 2 returns per pulse with an average accuracy of 55 mm RMSE in z
173 values at a 50 m range. The average survey altitude was set to 85 m and the average
174 approximate survey flight speed was 8.02 m/s.

175 The trajectory information was managed using PhoenixLiDAR System's Navlab to
176 improve system position and altitude. A precise post-processed trajectory was produced
177 through the combined integration of GNSS and IMU data captured by the LiDAR system.
178 The point cloud was derived in ETRS89 / UTM zone 30N (EPSG:25830) coordinate
179 system and categorized into ground, non-ground, and noise. Subsequently, a digital
180 elevation model (DEM), a digital surface model (DSM), and a canopy height model
181 (CHM) were computed.

182 The CHM was computed from the difference between the DEM and DSM at 3 cm
183 resolution. The DEM raster was obtained by selecting the ground points to determine the
184 ground level height above sea level. The DSM raster was created considering both
185 vegetation and ground points. The DEM and DSM were created using the LAS Dataset
186 to Raster tool in ArcMap version 10.7.1. The output raster's cell values were defined
187 through a binning approach, with an average assignment to each pixel and linear
188 interpolation for gap filling.

189 *2.3 Individual tree crown delineation*

190 Tree crown delineation was carried out manually. For the digitalization of 55 tree crowns,
 191 the true color RGB and false color (NIR-green-blue) multispectral orthomosaics were
 192 used, supported by the CHM calculated using the LiDAR data. This approach enabled
 193 precise identification of individual tree crowns, facilitating the subsequent analysis of
 194 multispectral and LiDAR metrics to study the presence of mistletoe within the canopy.
 195 This digitalization has been the basis for the computation of multispectral and LiDAR
 196 metrics for each individual tree. The manual delineation was accomplished using ArcMap
 197 software version 10.7.1.

198 *2.4. Metrics computation for individual trees*

199 The multispectral information was used to derive vegetation indices and textural features.
 200 Concretely, we computed sixteen vegetation indices at 3 cm resolution (see table 1). The
 201 use of this dataset of vegetation indices stems from the need to identify those that are
 202 most useful for the case study. They are key indicators used to assess health of plants
 203 (Thapa, 2013).

204 **Table 1.** Vegetation index calculated from spectral bands

Vegetation indices	Formula
NDVI (Normalized Difference Vegetation Index)	$NDVI = \frac{NIR - RED}{NIR + RED}$
GCI (Green Chlorophyll Index)	$GCI = \frac{NIR}{GREEN} - 1$
CIRE (Chlorophyll Index Red Edge)	$CIRE = \frac{NIR}{RE} - 1$
NDRE (Normalized Difference Red Edge)	$NDRE = \frac{NIR - RE}{NIR + RE}$
ETA (η)	$\eta = \frac{2(NIR^2 - RED^2) + 1.5NIR + 0.5RED}{NIR + RED + 0.5}$
GEMI (Global Environmental Monitoring Index)	$GEMI = \eta(1 - 0.25\eta) - \frac{RED - 0.125}{1 - RED}$
GNDVI (Green Normalized Difference Vegetation Index):	$GNDVI = \frac{NIR - GREEN}{NIR + GREEN}$
CVI (Chlorophyll Vegetation Index):	$CVI = \frac{NIR \times RED}{GREEN^2}$
SAVI (Soil-Adjusted Vegetation Index):	$SAVI = \frac{(1 + L)(NIR - RED)}{NIR + RED + L}$
MCARI (Modified Chlorophyll Absorption Ratio Index):	$MCARI = \left(\frac{RED - RE}{RE - GREEN}\right) \times NIR$
MSI (Moisture Stress Index):	$MSI = \frac{NIR}{RED}$
RI (Redness Index):	$RI = \frac{RED^2}{NIR \times BLUE}$
LCI (Leaf Chlorophyll Index):	$LCI = \frac{NIR - RE}{NIR + RED}$

CCCI (Canopy Chlorophyll Content Index): $CCCI = \frac{NIR - RE}{(NIR + RE)} / \frac{NIR - RED}{(NIR + RED)}$

RDVI (Renormalized Difference Vegetation Index): $RDVI = \frac{RED}{\sqrt{NIR + RED}}$

SQRT_SR (Square Root of Simple Ratio): $SQRT_SR = \sqrt{\frac{NIR}{RED}}$

205 The gray-level co-occurrence matrix (GLCM) was used to calculate textural features for
206 each multispectral band. Particularly, the contrast, homogeneity, dissimilarity, entropy,
207 second moment, mean, variance, correlation, and sum of averages features were selected
208 summing up a total of 45 metrics. Once the vegetation indices and textures were
209 computed, we derived the average value for each metric within the extent of the selected
210 tree crowns for subsequent individual analysis. Metric computation and processing were
211 carried out using the “glcm” and “raster” packages for R environment.

212 The extent of the digitized individual trees was used to clip the LiDAR-generated 3D
213 data, which was then normalized using LAStools to determine the height above ground
214 level. A set of structural metrics from the normalize LiDAR point clouds was derived
215 related to canopy height, height variability, and canopy density as described by Domingo
216 et al., (2024). A cutoff value of 3 m was applied to filter out ground and understory laser
217 returns before calculating the LiDAR metrics. The metrics were computed using FUSION
218 LDV v.4.50(McGaughey, 2023), lidRmetrics (Tompalski et al., 2024) and lasR (Næsset,
219 2004) in R environment.

220 Overall, a total of 167 variables were computed, which involved 16 spectral indices, 45
221 textural features, and 106 LiDAR structural metrics. These metrics served as based
222 information for the subsequent classification of infection levels (see 2.5).

223 *2.5. Classification of mistletoe infestation levels*

224 Firstly, we assessed the suitability of multispectral and LiDAR derived metrics for
225 mistletoe infestation levels discrimination. Initial analyses revealed that the data were not
226 normally distributed. Metrics were transformed to logarithmic and square root scales as
227 a feasible alternative to data normalization. Though, the Shapiro-Wilk test revealed that
228 the data distribution remained non-normal (p -value < 0.05). The nonparametric Kruskal-
229 Wallis test was used to identify which derived metrics shows statistically significant
230 differences between mistletoe infestation levels at the individual tree scale. Then, the non-
231 parametric Dunn’s post hoc test for multiple comparisons was applied to identify which
232 variables differentiated between pairs of trees with different levels of mistletoe infestation

233 and to determine the specific pairs distinguished by each variable (García-Galar et al.,
 234 2023; Hoffrén et al., 2023). While similar to the Kruskal-Wallis test, this method can
 235 identify the specific groups that shows statistically significant differences, making it
 236 useful for selecting suitable metrics that will be subsequently used for classification.

237 The performance of Random Forest (RF) and Support Vector Machine (SVM), two
 238 nonparametric machine learning classification algorithms, were evaluated for classifying
 239 trees according to their level of mistletoe infestation based on selected multispectral and
 240 LiDAR metrics. We tested various combinations of metrics for the models, including only
 241 multispectral data; only LiDAR data; or a combination of both datasets.

242 SVM was executed utilizing a radial kernel and parameterized with a cost of 200 and a
 243 gamma of 0.02. RF was tuned by implementing between 1,000 and 3,000 trees (ntrees)
 244 and between 1 and 2 variables in each node (mtry) according to Rodrigues and De la Riva
 245 (2014) and García-Galar et al. (2023), and the bias correction was applied. Models were
 246 computed in R using “e1071”, “MASS”, and “randomForest” packages. The dataset was
 247 divided into training and testing groups derived from a randomly selected sample of pixels
 248 to conduct the classification. The testing dataset was used to validate the models, executed
 249 by employing a 25% stratified random selection to cover the different mistletoe
 250 infestation levels. Validation was performed over 30 repetitions to obtain more robust
 251 results and mean performance values were calculated. To contrast and establish the best
 252 classification model, confusion matrices, user’s accuracy, producer’s accuracy and
 253 overall accuracy were assessed

254 3. Results

255 3.1. Selection of LiDAR and Multispectral metrics for mistletoe infestation levels 256 classification.

257 A total of 88 variables were found to be significant. Table 2 shows a selection of
 258 multispectral, LiDAR and textures variables with the highest chi-square values obtained
 259 after executing Kruskal-Wallis test.

260 **Table 2.** Results of the Kruskal-Wallis test.

Type	Metrics	Chi Square	p-value
Spectral indices	CIRE	44.70	***
	NDRE	44.47	***
	NDVI	39.61	***
	NIR mean	16.49	***

LiDAR	zsd	21.47	***
	zvar	21.47	***
	D1	18.69	***
	D0	18.06	***
	Elev L2	14.91	**
	% all returns above 3.00	12.48	**
	D9	8.61	*

261 *: p -value < 0.05; **: p -value < 0.01; ***: p -value < 0.001.

262 Vegetation indices CIRE, NDRE and NDVI presented the highest significant differences
 263 between mistletoe infestation levels while NIR mean textural metric showed a lower but
 264 significant value. *Viscum album* ssp. *austriacum* induces notable modifications in canopy
 265 structure, so LiDAR metrics related to canopy height metrics (moment 2 elevation),
 266 variability of canopy heights metrics (standard deviation and variation of the height), and
 267 canopy density metrics (D0, D1, D9 and percentage of all returns above 3 meters) were
 268 also significant.

269 After analyzing the metrics utilizing the Kruskal-Wallis test, the Dunn's test was applied.
 270 The results in table 3 shows the number of metrics that can differentiate each infestation
 271 level.

272 **Table 3.** Dunn's test results. Number of variables that can differentiate between infestation
 273 levels.

Groups	Number of variables
Level 1 – level 2	1
Level 1 – level 3	30
Level 1 – Level 4	69
Level 2 – level 3	0
Level 2 – level 4	17
Level 3 – level 4	20

274

275 **Table 4.** Dunn's test results.

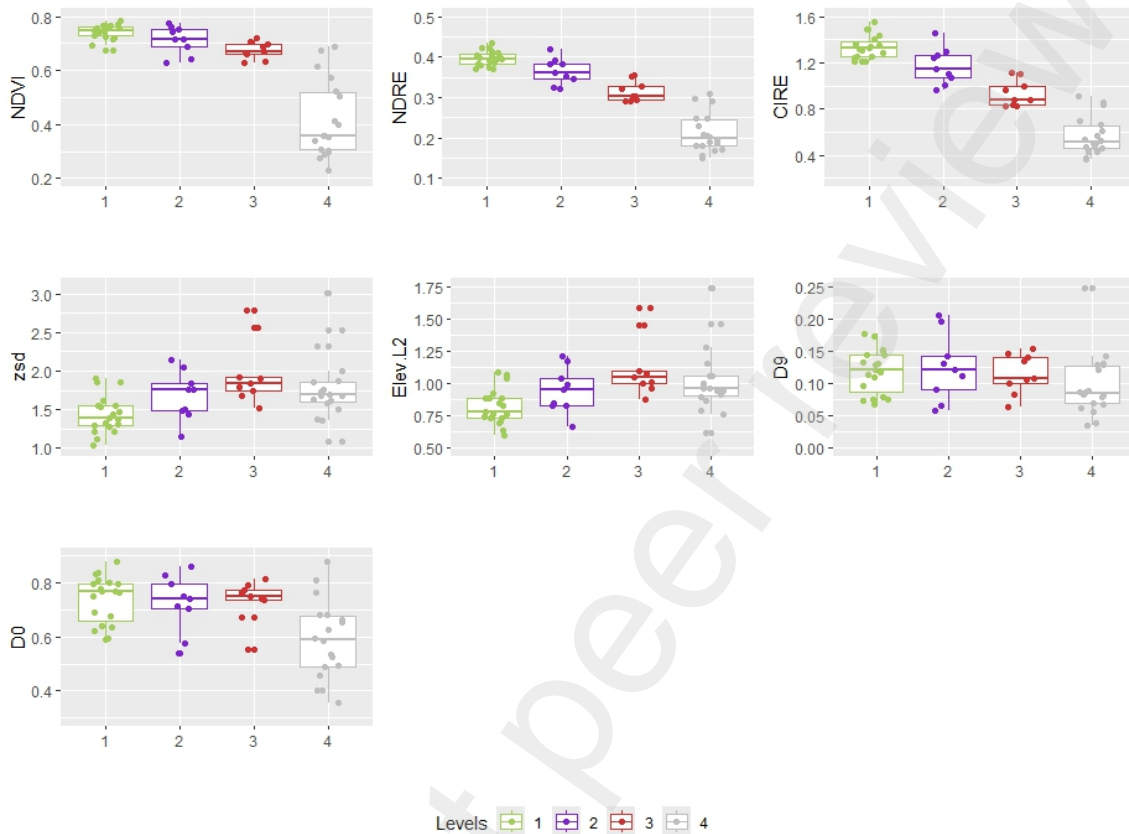
Variables	1-2	1-3	1-4	2-3	2-4	3-4
CIRE	0.8199 (NS)	0.0050 **	5.12E-10 ***	0.6725	0.0010 **	0.3294 (NS)
NDRE	0.8458 (NS)	0.0064 **	5.45E-10 ***	0.7344	0.0010 **	0.2868 (NS)
NDVI	1 (NS)	0.0392	6.05E-09 ***	1 (NS)	0.0009 ***	0.1563 (NS)
zsd	0.2923 (NS)	0.0044 **	0.0001 ***	1 (NS)	0.8502 (NS)	1 (NS)
zvar	0.2923 (NS)	0.0044 **	0.0001 ***	1 (NS)	0.8502 (NS)	1 (NS)

D1	1 (NS)	1 (NS)	0.0004 ***	1 (NS)	0.0650(NS)	0.0158 *
D0	1 (NS)	1 (NS)	0.0007 ***	1 (NS)	0.0394 *	0.0158 *
NIR mean	1 (NS)	1 (NS)	0.0035 *	1 (NS)	0.1819 (NS)	0.0048 *
	0.5927	0.0024 **	0.0287 *	0.6335	1 (NS)	1 (NS)
Elev L2	(NS)			(NS)		
% all returns above 3.00	1 (NS)	1 (NS)	0.0059 **	1 (NS)	0.1597 (NS)	0.1280 (NS)
	1 (NS)	1 (NS)	0.0373 *	1 (NS)	0.2646 (NS)	0.5458 (NS)
D9						

276 *: p -value < 0.05; **: p -value < 0.01; ***: p -value < 0.001; NS: non-significant.

277 Overall, 13 vegetation indices were able to discern between 3 pairs of mistletoe
278 infestation levels. Additionally, a LiDAR-derived variable, D0, was also able to discern
279 3 pairs. A total of 26 metrics were able to discern between 2 pairs. These include LiDAR-
280 based variables associated with the variability of canopy heights, the distribution of
281 canopy heights, and canopy density, such standard deviation of elevation values, variation
282 of elevation values, 1st percentile representing the value below which 1% of the data is
283 found, and interquartile distance, respectively. Included in these are metrics derived from
284 multispectral bands, such as the vegetation indices ETA, GEMI, and CCCI, as well as
285 those related to textures of the NIR and blue bands.

286 Figure 2 presents the selected variables that demonstrate statistically significant
 287 differences in Dunn's test between mistletoe infestation levels. The greatest differences
 288 were found between infestation Level 1 respect to Level 4



289 **Figure 2.** Variables with statistically significant differences in Dunn's test between mistletoe
 290 infestation levels.

291

292 *3.2 Classification of Mistletoe infestation levels classification models.*

293 Mistletoe affects spectral response and structure of the canopy. Vegetation indices were
 294 the most important variables for infested trees classification using radial kernel in the
 295 SVM algorithm. Using two vegetation indices (CIRE and NDVI) an overall accuracy of
 296 0.83 was achieved by applying SVM, while combining a spectral metric (NDRE) and a
 297 LiDAR metric (D0) resulted in 0.82 accuracy with SVM. Using only LiDAR variables,
 298 we obtained an accuracy of 0.64 with SVM and RF. Table A1 lists the selected metrics
 299 from various groups and sensors, and table 5 presents the results metrics combinations
 300 and classification methods.

301

302

303 **Table 5.** Support Vector Machine (SVM) and Random Forest (RF) Classification methods.

Types of metrics	Metrics	Methods	Fitting phase	Validation
Spectral indices	CIRE+NDVI	SVM	0.823	0.833
	CIRE+NDVI	RF	1	0.733
Spectral indices + LiDAR	NDRE + D0	SVM	0.817	0.817
	NDVI + Elev.L2	RF	1	0.743
LiDAR	zsd + % all ret. above 3.00+ D9	SVM	0.753	0.643
	zsd + D0 + D9	RF	1	0.645

304

305 The parameters chosen for the SVM were cost = 200, gamma = 0.02 and a radial basis
 306 function kernel. For RF, ntrees = 2000, mtry = 1 with bias correction. Using spectral
 307 metrics results in a 15.28% improvement for SVM compared to RF. LiDAR metrics show
 308 no difference in performance between the two algorithms. Combining both LiDAR and
 309 multispectral metrics, leads to a 9.46% improvement for SVM compared to RF.

310 Overall, the models shown in Tables 6 to 8, which incorporate spectral, LiDAR, and
 311 combined spectral and LiDAR data, exhibited lower accuracy at levels 2 and 3. Errors
 312 between close infestation levels were to be expected. The boxplots in Figure 2 show the
 313 metric similarities, especially between infestation levels 1, 2, and 3.

314 **Table 6.** Confusion matrix of the best SVM model with spectral indices metrics.

CIRE+NDVI	Level 1	Level 2	Level 3	Level 4	User's accuracy
Level 1	19	4	0	0	82.61%
Level 2	0	4	2	0	66.67%
Level 3	0	1	7	3	63.64%
Level 4	0	0	0	15	100.00%
Prod.'s accuracy	100.00%	44.44%	77.78%	83.33%	

315

316 **Table 7.** Confusion matrix of the best SVM model with LiDAR metrics.

zsd+% all ret. above 3.00+ D9	Level 1	Level 2	Level 3	Level 4	User's accuracy
Level 1	17	3	0	2	77.27%
Level 2	0	5	2	2	55.56%
Level 3	2	0	6	1	66.67%

Level 4	0	1	1	13	86.67%
Prod.'s accuracy	89.47%	55.56%	66.67%	72.22%	

317

318 **Table 8.** Confusion matrix of the best SVM model with spectral indices + LiDAR metrics

NDRE + D0	Level 1	Level 2	Level 3	Level 4	User's accuracy
Level 1	19	4	0	0	82.61%
Level 2	0	5	3	0	62.50%
Level 3	0	0	6	2	75.00%
Level 4	0	0	0	16	100.00%
Prod.'s accuracy	100%	55.56%	66.67%	88.89%	

319

320 The three classification models (Tables 6 to 8) show greater confusion when classifying
 321 trees with infestation levels 2 and 3. The producer's accuracy for level 2 ranges from 44%
 322 to 55%, while for level 3 it ranges from 66% to 78%. In contrast, level 1 shows an
 323 accuracy of 90% to 100%, and level 4 ranges from 72% to 89%. The user's accuracy
 324 exceeds 80% in all three models at level 1, ranges from 62% to 66% at level 2, from 63%
 325 to 75% at level 3, and from 86% to 100% at level 4.

326 **4. Discussion**

327 The effect of mistletoe in pine forests leads to changes in crown vigor and canopy
 328 structure. Mistletoe host experiences a reduction in its water and nutritional supply,
 329 causing a progressive atrophy that develops from the implantation site (Hernández-
 330 Alonso et al., 2001). This can lead to crown transparency, dead branches, and needle
 331 discoloration (Dobbertin, 2005), potentially causing changes in the spectral response and
 332 structure of the canopies. Mistletoe infestations lead to reductions in chlorophyll index
 333 red edge (CIRE), with a 57.9% decrease between infestation level 1 and level 4, and a
 334 30.1% decrease between level 1 and level 3. The standard deviation of the LiDAR point
 335 clouds of trees at level 3 is 37.76% higher compared to the standard deviation at level 1.

336 Traditional control methods for mistletoe have been costly, highlighting the need for more
 337 efficient alternatives for monitoring infestations in Mediterranean pine forests. Standard
 338 field methods have limitations, and it is necessary to call for new approaches to assess
 339 tree health based on remote sensing, concretely related to mistletoe affection (Ančić et
 340 al., 2014). Our study offers a new perspective for forest monitoring by examining
 341 individual trees and differentiating infestation levels at an individual scale. Damage
 342 severity classification is more relevant for decision-making than mere damage detection,

343 as damage quantification is required in forest pest management (Rullán-Silva et al.,
344 2015). The utilization of UAV platforms opens new ways for analyzing individual tree
345 scales with a three-dimensional perspective on canopy changes (Domingo et al., 2024b).

346 Spectral and LiDAR UAV derived data, complemented by field sampling, enable the
347 discrimination of infestation levels. Statistically significance differences are identified,
348 and the presence of mistletoe resulted in modifications to canopy morphology and
349 spectral crown response. NDRE, CIRE and NDVI are the most important and significant
350 metrics. The spectral response of infested and non-infested trees differs, particularly
351 across four levels of mistletoe infestation. Spectral signatures reveal insights into the state
352 and structure of both leaves and canopy (Huete, 2012). Damage to host trees causes
353 defoliation and reduced growth (Galiano et al., 2010) eventually leading to tree dead
354 (Reid et al., 1994). Significant reductions in CIRE, NDRE, and NDVI vegetation indices
355 values are observed in the upper canopy due to mistletoe infestation. These variables
356 consider the red, red-edge, and NIR bands. RGB and multispectral UAV data were used
357 by Mejia-Zuluaga et al. (2022), Mischyszyn & Wezyk (2022), León-Bañuelos et al.
358 (2020), and Miraki et al. (2021) to detect different mistletoes species using visible and
359 NIR bands or vegetation indices. Maes et al. (2018) employed UAV based infrared
360 thermography, showing that the surface temperature of the eucalypt foliage of infested
361 trees was notably higher. Our study also introduces key metrics derived from UAV
362 LiDAR, a data source previously employed by researchers such as Barbosa et al. (2016).
363 However, we leverage LiDAR metrics to distinguish between different levels of
364 infestation, marking a novel application of this technology. Height variability, height
365 distribution and canopy cover density have been the metrics used in classification models.
366 Differences in infestation levels are associated with a more variable and heterogeneous
367 canopy.

368 The categorization into four levels of mistletoe infection allows us to see which classes
369 are difficult to differentiate. Significant differences are observed between level 1 and level
370 4, and the most notable confusions occur between the intermediate levels, as shown
371 confusion matrices. This aspect could be explained by the similarity in the structure of
372 trees with similar levels of infestation. LiDAR information is very useful in this context,
373 providing a three-dimensional perspective of pine trees infested by mistletoe. Tree-level
374 metrics, such as canopy heights, variability, and density capture morphological
375 transformations across different infestation degrees. Multispectral data indicate tree

376 health and mistletoe presence in the upper canopy, with spectral responses varying
377 according to infestation severity. Other forest pests, such as the pine processionary moth,
378 have been shown to alter canopy cover and to cause reductions in the upper canopy, as
379 reported by Domingo et al. (2024) using leaf area index (LAI) and mean leaf area density
380 metrics derived from LiDAR.

381 Miraki et al. (2021) achieve reliable performance using RF to differentiate between two
382 categories (infested and non-infested trees) with both manual and automatic crown
383 segmentation of photogrammetry-derived data in leaf-on and leafless conditions. The
384 overall accuracy is 0.87 for manual segmentation under leaf-off conditions, and 0.76 for
385 the combined leaf-off and leaf-on situations. Barbosa et al. (2016) demonstrate good
386 performance using SVM with a radial function kernel, based on consistent spectral. This
387 approach achieves an accuracy of 86% in classifying two classes: presence and absence
388 of mistletoe. Barbosa et al., using LiDAR data, found that the landscape structure
389 influences the presence of mistletoe, with isolated host trees exhibit twice the infestation
390 load compared to those at the core of forest fragments. In our research we use SVM and
391 RF algorithms to distinguish four infestation levels. When using spectral metrics with
392 SVM (OA = 0.83), although the combination of both sensors also demonstrates
393 significant accuracy (OA = 0.82). RF is not as accurate as SVM in this case, although it
394 achieves optimal results for classifying infestation levels using spectral metrics (OA =
395 0.73) and by combining spectral and LiDAR data (OA = 0.74). Using only three LiDAR
396 metrics, both algorithms achieved the same accuracy (0.64).

397 Our study demonstrates the potential of MS and LiDAR UAV data in delineating
398 structural and spectral differences in the crowns infested from *Viscum album* ssp.
399 *austriacum* within a Mediterranean forest dominated by Scots pine. Statistical differences
400 are identified through Kruskal-Wallis and Dunn's test also identified significant
401 differences. These variables are susceptible to morphological and spectral changes in the
402 crowns of host trees according to four infestation levels. His methodology could be
403 applied in future research to wilder areas by developing statistical models to improve
404 forest management. Our outcomes provide insights for similar applications by presenting
405 relevant MS and LiDAR metrics that can be developed into prediction models for other
406 areas. According to Maes et al. (2018) thermal imagery could enable the analysis between
407 infested and non-infested mistletoe trees. Another potential line of research could be to
408 further develop the work initiated by Barbosa et al. (2016) focusing on the landscape

409 characterization in forest stands infested by this hemiparasitic plant, distinguishing
410 between isolated trees, forest edge and forest interior. Combining remote sensing
411 technologies, both active and passive, would be a particularly valuable approach. Satellite
412 imagery could offer the advantage of covering large areas, facilitating the identification
413 of infestation patterns at a regional scale. By integrating these satellite images with UAV
414 data, which provides detailed information on canopy structure and spectral response using
415 multispectral and LiDAR sensors, detection, monitoring, and management strategies for
416 mistletoe in forest ecosystems are significantly enhanced. This combination of tools
417 would enable a more comprehensive and accurate understanding of mistletoe infestation,
418 optimizing its management and control.

419 **5. Conclusion**

420 This research evaluated the potential of combining UAV-derived multispectral imagery
421 and LiDAR point cloud to determine *Viscum album* ssp. *austriacum* affection levels. Our
422 approach demonstrates their value for detecting and characterizing vegetation vigor and
423 morphological changes in up to four levels of mistletoe infestation in Mediterranean Scots
424 pine forest. The most accurate mistletoe infestation classification model was developed
425 using the radial kernel SVM, incorporating two spectral variables: CIRE, and NDVI. The
426 model classification obtained an overall accuracy of 0.83 after validation. LiDAR point
427 cloud derived metrics with RF model achieved a global accuracy of 0.64, which included:
428 standard deviation of elevation values, D0 and D9. Combining both sensor active and
429 passive, classification model with SVM radial kernel achieved an overall accuracy of
430 0.82, with NDRE and D0. UAV data supports the monitoring of forest management
431 regarding a hemiparasitic plant that currently threatens Mediterranean forests in a context
432 of global change.

433

434 **Funding**

435 This research was funded by Gobierno de Aragón-Fondo de Inversiones de Teruel (FITE)
436 and Gobierno de España, grant project FITE-2021-DRUIDA and by Gobierno de Aragón
437 research groups S51_23R and S74_23R.

438 **Acknowledgements**

439 We would like to thank the members of the Forest Management Unit-Provincial Service
440 of Teruel for their valuable assistance during the field work.

441 Appendix A

442 **Table A1.** Final metrics included in the models

Group of variables	Variables	Description
Height distribution	Elev.L2	L moment 2 elevation
Height variability	zsd	Standard deviation of elevation values
	zvar	Variation of elevation values
Canopy cover density	D0, D1, D9	% of all returns in 10 equally distributed vertical layers derived by separating the height between the 95 th percentile of the height distribution and the 3 m threshold
	% all ret. above 3.00	Percentage of all returns above 3.00
Spectral indices	NDRE	Normalized Difference Red Edge
	CIRE	Index Chlorophyll Index - Red-Edge
	NDVI	Normalized Difference Vegetation Index
Textural features	NIR mean	Mean of the NIR band

443

444 **Table A2.** Number of groups that can differentiate each variable

Variables	Number of groups
NDVI, GCI, CIRE, NDRE, GNDVI, CVI, SAVI, MCARI, MSI, RI, LCI, RDVI, SQRT_SR, D0	3
zvar, zsd, zq1, ziqr, zMADmean, zMADmedian, zpcum1, L2, elev.stddev, elev_variance, elev.cv, elev.AAD, elev.MAD.mode, elev.L2, elev.L.CV, Hsd, D1, D2, eta, GEMI, CCCI, NIR_mean, NIR_SA, Blue_ASM, Blue_mean, Blue SA	2
zmin, zcv, zq5, pzabove2, zentropy, zpcum2, zpcum3, zpcum4, Lcoefvar, lad_max, lad_mean, lad_sum, pz_0.15.2, pz_5.6, pz_8.5.10, Return.2.count, Elev.IQ, Elev.MAD.median, Percentage.first.returns.above.3.00, Percentage.all.returns.above.3.00, Percentage.all.returns.above.mean, Hcv, D3, D4, D5, D6, D7, D8, D9, NIR_homogeneity, NIR_ASM, NIR_entropy, NIR_correlation, RE_homogeneity, RE_ASM, RE_entropy, RE_mean, RE_SA, Red_ASM, Red_mean, Red_correlation, Red SA, Blue correlation	1

445

446

447

448

449

450

451

452 REFERENCES

- 453 Allen, C.D., Macalady, A.K., Chenchouni, H., Bachelet, D., McDowell, N., Vennetier,
454 M., Kitzberger, T., Rigling, A., Breshears, D.D., Hogg, E.H. (Ted), Gonzalez, P.,
455 Fensham, R., Zhang, Z., Castro, J., Demidova, N., Lim, J.H., Allard, G., Running, S.W.,
456 Semerci, A., Cobb, N., 2010. A global overview of drought and heat-induced tree
457 mortality reveals emerging climate change risks for forests. *For. Ecol. Manag.* 259, 660–
458 684. <https://doi.org/10.1016/j.foreco.2009.09.001>
- 459 Ančić, M., Pernar, R., Bajić, M., Seletković, A., Kolić, J., 2014. Detecting mistletoe
460 infestation on silver fir using hyperspectral images. *IForest* 7, 85–91.
461 <https://doi.org/10.3832/ifor1035-006>
- 462 Barbosa, J.M., Sebastián-González, E., Asner, G.P., Knapp, D.E., Anderson, C., Martin,
463 R.E., Dirzo, R., 2016. Hemiparasite–host plant interactions in a fragmented landscape
464 assessed via imaging spectroscopy and LiDAR. *Ecol. Appl.* 26, 55–66.
465 <https://doi.org/10.1890/14-2429>
- 466 Dobbertin, M., 2006. Pine mistletoe (*Viscum album* ssp. *austriacum*) contributes to Scots
467 pine (*Pinus sylvestris*) mortality in the Rhone valley of Switzerland. *For. Pathol.* 36, 309–
468 322. <https://doi.org/10.1111/J.1439-0329.2006.00457.X>
- 469 Dobbertin, M., 2005. Tree growth as indicator of tree vitality and of tree reaction to
470 environmental stress: a review. *Eur. J. For. Res.* 124(4), 319–333.
471 <https://doi.org/10.1007/s10342-005-0085-3>
- 472 Domingo, D., Gómez, C., Mauro, F., Houdas, H., Sangüesa-Barreda, G., Rodríguez-
473 Puerta, F., 2024. Canopy Structural Changes in Black Pine Trees Affected by Pine
474 Processionary Moth Using Drone-Derived Data. *Drones*, 8(3), 75.
475 <https://doi.org/10.3390/drones8030075>
- 476 Ecke, S., Dempewolf, J., Frey, J., Schwaller, A., Endres, E., Klemmt, H.J., Tiede, D.,
477 Seifert, T., 2022. UAV-Based Forest Health Monitoring: A Systematic Review. *Remote*
478 *Sens.* 14(13), 3205. <https://doi.org/10.3390/rs14133205>
- 479 Galiano, L., Martínez-Vilalta, J., Lloret, F., 2010. Drought-Induced Multifactor Decline
480 of Scots Pine in the Pyrenees and Potential Vegetation Change by the Expansion of Co-
481 occurring Oak Species. *Ecosystems* 13, 978–991. <https://doi.org/10.1007/s10021-010->
482 [9368-8](https://doi.org/10.1007/s10021-010-9368-8)

483 García-Galar, A., Lamelas, M.T., Domingo, D., 2023. Assessment of Oak Groves
484 Conservation Statuses in Natura 2000 Sacs with Single Photon Lidar and Sentinel-2 Data.
485 Remote Sens. 15(3), 710. <https://doi.org/10.3390/rs15030710>

486 Glatzel, G., Geils, B.W., 2009. Mistletoe ecophysiology: Host-parasite interactions.
487 Botany, 87(1), 10-15. <https://doi.org/10.1139/B08-096>

488 Gobierno de Aragón, Departamento de Desarrollo Rural y Sostenibilidad, Dirección
489 General de Gestión Forestal, 2020. Redes de evaluación fitosanitaria en las masas
490 forestales de Aragón. Manual de campo. 25 pp. Gobierno de Aragón.

491 Guimarães, N., Pádua, L., Marques, P., Silva, N., Peres, E., Sousa, J.J., 2020. Forestry
492 Remote Sensing from Unmanned Aerial Vehicles: A Review Focusing on the Data,
493 Processing and Potentialities. Remote Sens. 2020, 12(6), 1046.
494 <https://doi.org/10.3390/RS12061046>

495 Hernández-Alonso, R., Martín Bernal, E., Cañada Martín, J.F., Gil Pelegrín, E., Zorrilla
496 Alcaine, F., Pérez Fortea, V., Delgado Soriano, J., Ibarra Ibañez, N., 2001. El muérdago.
497 Gobierno de Aragón, Dirección General del Medio Natural, Servicio de Estudios,
498 Coordinación y Defensa Contra Incendios Forestales. Available at:
499 https://bibliotecavirtual.aragon.es/i18n/catalogo_imagenes/grupo.cmd?path=3714805
500 (accessed on 1 December 2024)

501 Hernández-Jiménez, Á., 2020. Evolución de la afección del muérdago (*Viscum album*
502 subsp. *austriacum*) en pinares de la provincia de Zaragoza. Líneas de actuación. Sección
503 Provincial de Zaragoza, Departamento de Agricultura, Ganadería y Medio Ambiente,
504 Gobierno de Aragón. Available at:
505 [https://distritoforestal.es/images/AFECCI%C3%93N_DEL_MU%C3%89RDAGO_EN](https://distritoforestal.es/images/AFECCI%C3%93N_DEL_MU%C3%89RDAGO_EN_LA_PROVINCIA_DE_ZARAGOZA_baja_res.pdf)
506 [LA_PROVINCIA_DE_ZARAGOZA_baja_res.pdf](https://distritoforestal.es/images/AFECCI%C3%93N_DEL_MU%C3%89RDAGO_EN_LA_PROVINCIA_DE_ZARAGOZA_baja_res.pdf)

507 Hoffrén, R., Lamelas, M.T., de la Riva, J., 2023. UAV-derived photogrammetric point
508 clouds and multispectral indices for fuel estimation in Mediterranean forests. Remote
509 Sens. Appl. 31. <https://doi.org/10.1016/j.rsase.2023.100997> (accessed on 1 December
510 2024)

511 Huete, A.R., 2012. Vegetation Indices, Remote Sensing and Forest Monitoring. Geogr.
512 Compass 6, 513–532. <https://doi.org/10.1111/J.1749-8198.2012.00507.X>

513 León-Bañuelos, L.A., Endara-Agramont, A.R., Gómez-Demetrio, W., Martínez-García,
514 C.G., Gabino Nava-Bernal, E., 2020. Identification of *Arceuthobium globosum* using
515 unmanned aerial vehicle images in a high mountain forest of central Mexico. J. For. Res.
516 31, 1759–1771. <https://doi.org/10.1007/s11676-019-00954-5>

517 Maes, W.H., Huete, A.R., Avino, M., Boer, M.M., Dehaan, R., Pendall, E., Griebel, A.,
518 Steppe, K., 2018. Can UAV-based infrared thermography be used to study plant-parasite
519 interactions between mistletoe and Eucalypt trees? Remote Sens. 10(12) 2062.
520 <https://doi.org/10.3390/rs10122062>

521 Mathiasen, R.L., Nickrent, D.L., Shaw, D.C., Watson, D.M., 2008. Mistletoes: Pathology,
522 systematics, ecology, and management. Plant Dis. 92(7), 988-1006.
523 <https://doi.org/10.1094/PDIS-92-7-0988>

524 Mcgaughey, R.J., 2023. FUSION/LDV: Software for LIDAR Data Analysis and
525 Visualization V.4.21. USDA Forest Service. Washington DC, USA.

526 Mejia-Zuluaga, P.A., Dozal, L., Valdiviezo-N., J.C., 2022. Genetic Programming
527 Approach for the Detection of Mistletoe Based on UAV Multispectral Imagery in the
528 Conservation Area of Mexico City. Remote Sens. 14(3), 801.
529 <https://doi.org/10.3390/rs14030801>

530 Miraki, M., Sohrabi, H., Fatehi, P., Kneubuehler, M., 2021. Detection of mistletoe
531 infected trees using UAV high spatial resolution images. J. Plant Dis. Prot. 128, 1679–
532 1689. <https://doi.org/10.1007/s41348-021-00502-6>

533 Missarov, A., Krasylenko, Y., Krůček, M., Slavik, M., Nutfullin, B., Král, K., 2021.
534 Drone LiDAR remote sensing for mistletoe recognition and monitoring. Proceedings of
535 the SilviLaser Conference 2021, 300–302. <https://doi.org/10.34726/wim.2016>

536 Missarov, A., Sosnovsky, Y., Rydlo, K., Brovkina, O., Maes, W.H., Král, K., Krůček,
537 M., Krasylenko, Y., 2024. Vertical botany: airborne remote sensing as an emerging tool
538 for mistletoe research. Botany, 102(3), 58–71. <https://doi.org/10.1139/cjb-2023-0049>

539 Mischczynyn, J., Wezyk, P., 2022. Detection of mistletoe (*Viscum album* ssp. *austriacum*
540 L.) in Scots pine stands using the MS and RGB UAV high resolution orthophoto. Poster.
541 Available at: https://www.researchgate.net/figure/Poster_fig1_375592451

542 Næsset, E., 2004. Practical large-scale forest stand inventory using a small-footprint
543 airborne scanning laser. Scand J. For. Res. 19(2), 164–179.
544 <https://doi.org/10.1080/02827580310019257>

545 Reid, N., Yan, Z., Fittler, J., 1994. Impact of mistletoes (*Amyema miquelii*) on host
546 (*Eucalyptus blakelyi* and *Eucalyptus melliodora*) survival and growth in temperate
547 Australia. For. Ecol. Manag. 70(1-3), 55–65. [https://doi.org/10.1016/0378-](https://doi.org/10.1016/0378-1127(94)90074-4)
548 [1127\(94\)90074-4](https://doi.org/10.1016/0378-1127(94)90074-4)

549 Rullán-Silva, C., Olthoff, A.E., Pando, V., Pajares, J.A., Delgado, J.A., 2015. Remote
550 monitoring of defoliation by the beech leaf-mining weevil *Rhynchaenus fagi* in northern
551 Spain. For. Ecol. Manag. 347, 200–208. <https://doi.org/10.1016/j.foreco.2015.03.005>

552 Sabrina, F., Sohail, S., Thakur, S., Azad, S., Wasimi, S., 2020. Use of Deep Learning
553 Approach on UAV imagery to Detect Mistletoe Infestation. IEEE Region 10 Symposium,
554 TENSYPMP 2020., Dhaka, 556–559.
555 <https://doi.org/10.1109/TENSYPMP50017.2020.9230971>

556 Sancho-Knapik, D., Sanz, M.Á., Peguero-Pina, J.J., Niinemets, Ü., Gil-Pelegrín, E., 2017.
557 Changes of secondary metabolites in *Pinus sylvestris* L. needles under increasing soil
558 water deficit. Ann. For. Sci. 74, 24. <https://doi.org/10.1007/s13595-017-0620-7>

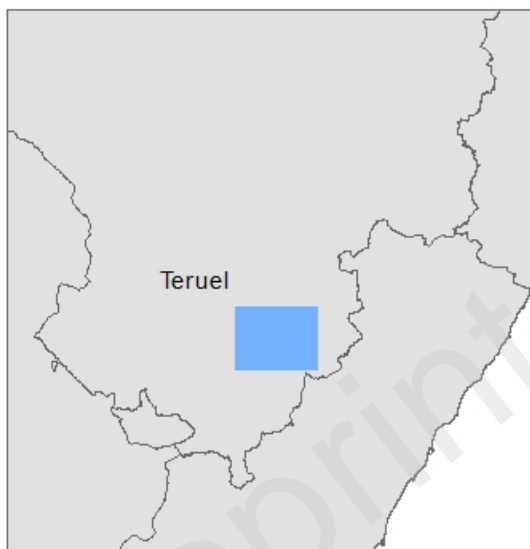
559 Sangüesa-Barreda, G., Camarero, J.J., Pironon, S., Gazol, A., Peguero-Pina, J.J., Gil-
560 Pelegrín, E., 2018. Delineating limits: Confronting predicted climatic suitability to field
561 performance in mistletoe populations. J. Ecol. 106, 2218–2229.
562 <https://doi.org/10.1111/1365-2745.12968>

563 Sangüesa-Barreda, G., Linares, J.C., Julio Camarero, J., 2013. Drought and mistletoe
564 reduce growth and water-use efficiency of Scots pine. For. Ecol. Manage 296, 64–73.
565 <https://doi.org/10.1016/J.FORECO.2013.01.028>

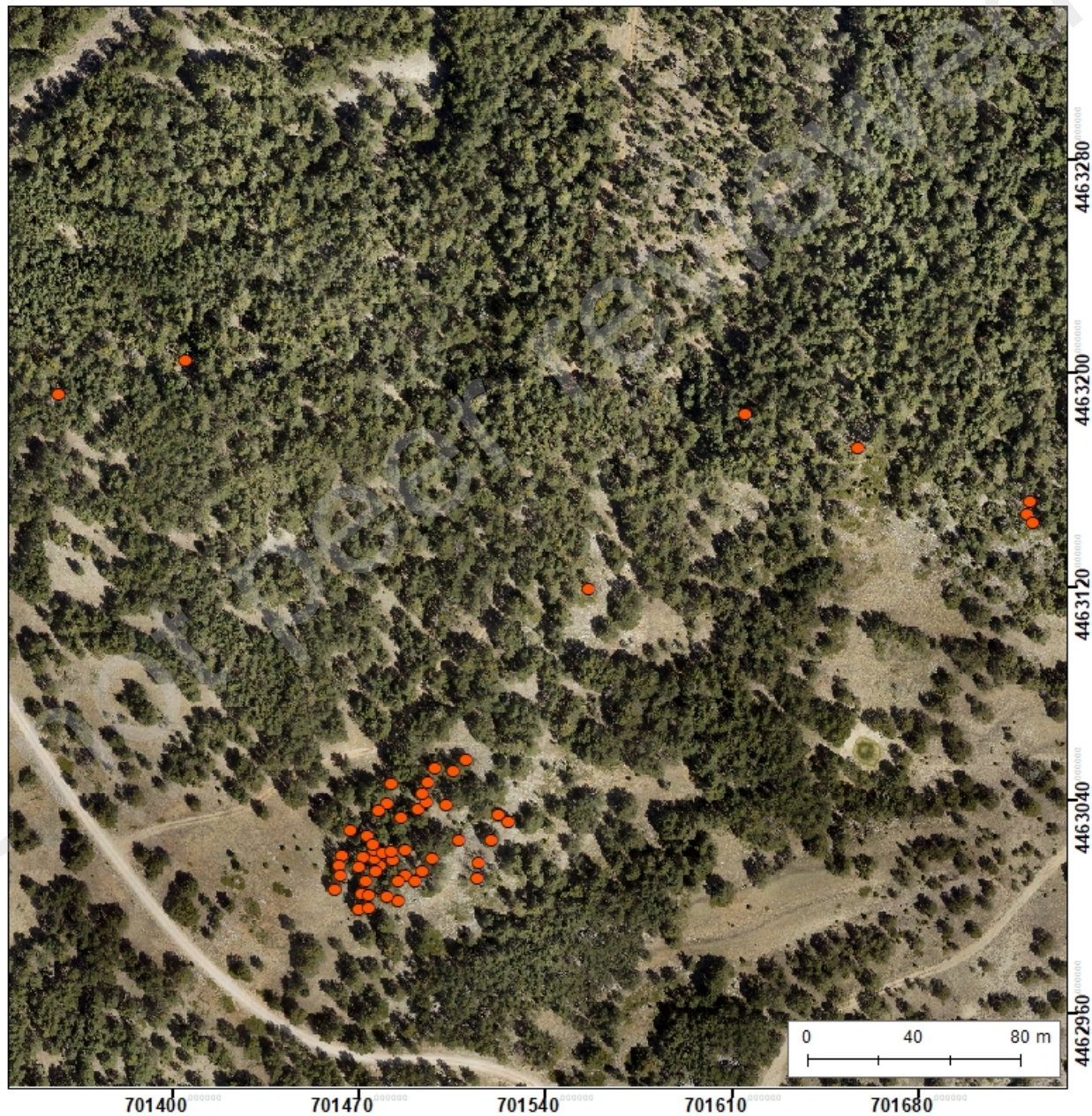
566 Senf, C., Pflugmacher, D., Zhiqiang, Y., Sebald, J., Knorn, J., Neumann, M., Hostert, P.,
567 Seidl, R., 2018. Canopy mortality has doubled in Europe’s temperate forests over the last
568 three decades. Nat. Commun 9, 4978. <https://doi.org/10.1038/s41467-018-07539-6>

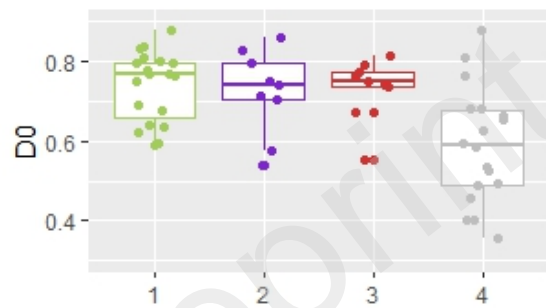
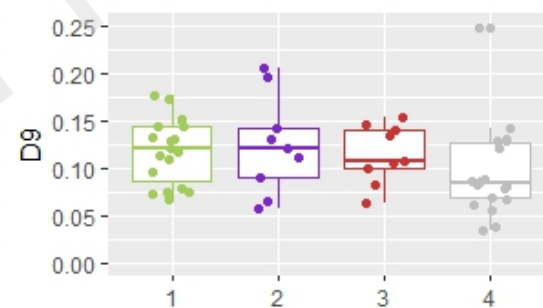
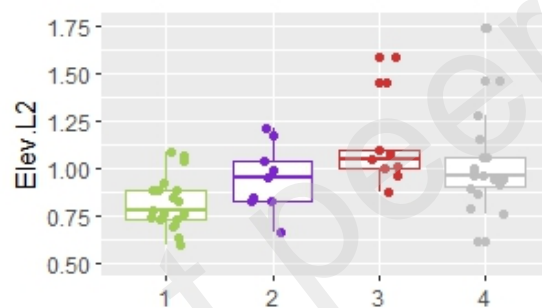
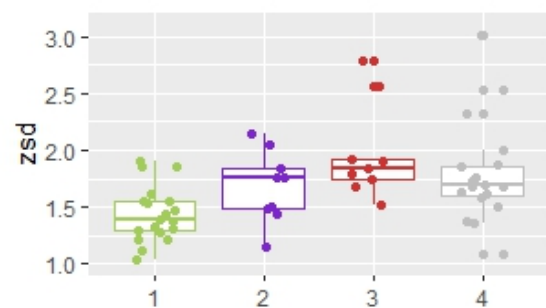
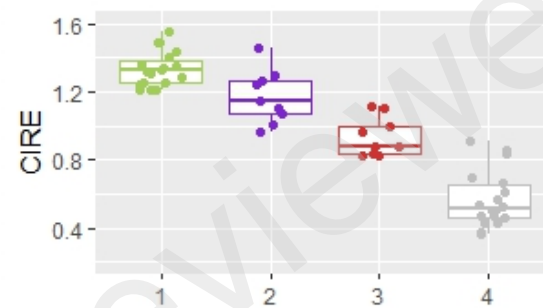
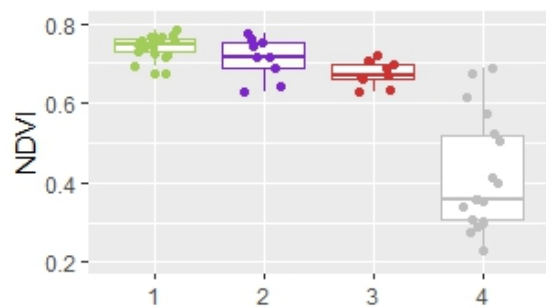
569 Thapa, S., 2013. Detection and mapping of incidence of *Viscum album* in *Pinus sylvestris*
570 forest of southern Alpe using satellite and airborne imagery. [Master’s thesis]. University
571 of Twente. Available at: <http://essay.utwente.nl/84727/1/thapa.pdf>

- 572 Tompalski, P.; Roussel, J.; Woods, M.; Hambrecht, L. Available at:
573 <https://github.com/ptompalski/lidRmetrics> (accessed on 3 December 2024).
- 574 Torres, P., Rodes-blanco, M., Viana-soto, A., Nieto, H., García, M., 2021. The Role of
575 Remote Sensing for the Assessment and Monitoring of Forest Health: A Systematic
576 Evidence Synthesis. *Forests* 12, 1134. <https://doi.org/10.3390/F12081134>
- 577 Tymińska-Czabańska, L., Janiec, P., Hawryło, P., Ślopek, J., Zielonka, A., Netzel, P.,
578 Janczyk, D., Socha, J., 2024. Modeling the effect of stand and site characteristics on the
579 probability of mistletoe infestation in Scots pine stands using remote sensing data. *For.*
580 *Ecosyst.* 11, 100191. <https://doi.org/10.1016/J.FECS.2024.100191>
- 581 Zuber, D., 2004. Biological Flora of Central Europe Biological flora of Central Europe:
582 *Viscum album* L. *Flora* 199, 181-203. <https://doi.org/10.1078/0367-2530-00147>



- Sampled trees
- Sierra de Gúdar





Levels  1  2  3  4

Table 1. Vegetation index calculated from spectral bands

Vegetation index	Formula
NDVI (Normalized Difference Vegetation Index)	$NDVI = \frac{NIR - RED}{NIR + RED}$
GCI (Green Chlorophyll Index)	$GCI = \frac{NIR}{GREEN} - 1$
CIRE (Chlorophyll Index Red Edge)	$CIRE = \frac{NIR}{RE} - 1$
NDRE (Normalized Difference Red Edge)	$NDRE = \frac{NIR - RE}{NIR + RE}$
ETA (η)	$\eta = \frac{2(NIR^2 - RED^2) + 1.5NIR + 0.5RED}{NIR + RED + 0.5}$
GEMI (Global Environmental Monitoring Index)	$GEMI = \eta(1 - 0.25\eta) - \frac{RED - 0.125}{1 - RED}$
GNDVI (Green Normalized Difference Vegetation Index):	$GNDVI = \frac{NIR - GREEN}{NIR + GREEN}$
CVI (Chlorophyll Vegetation Index):	$CVI = \frac{NIR \times RED}{GREEN^2}$
SAVI (Soil-Adjusted Vegetation Index):	$SAVI = \frac{(1 + L)(NIR - RED)}{NIR + RED + L}$
MCARI (Modified Chlorophyll Absorption Ratio Index):	$MCARI = \left(\frac{RED - RE}{RE - GREEN}\right) \times NIR$
MSI (Moisture Stress Index):	$MSI = \frac{NIR}{RED}$
RI (Redness Index):	$RI = \frac{RED^2}{NIR \times BLUE}$
LCI (Leaf Chlorophyll Index):	$LCI = \frac{NIR - RE}{NIR + RED}$
CCCI (Canopy Chlorophyll Content Index):	$CCCI = \frac{NIR - RE}{(NIR + RE)} / \frac{NIR - RED}{(NIR + RED)}$
RDVI (Renormalized Difference Vegetation Index):	$RDVI = \frac{RED}{\sqrt{NIR + RED}}$
SQRT_SR (Square Root of Simple Ratio):	$SQRT_SR = \sqrt{\frac{NIR}{RED}}$

Table 2. Results of the Kruskal-Wallis test.

Type	Metrics	Chi Square	p-value
Spectral indices	CIRE	44.70	***
	NDRE	44.47	***
	NDVI	39.61	***
	NIR mean	16.49	***
	zsd	21.47	***
	zvar	21.47	***
LiDAR	D1	18.69	***
	D0	18.06	***
	Elev L2	14.91	**
	% all returns above 3.00	12.48	**

*: p -value < 0.05; **: p -value < 0.01; ***: p -value < 0.001.

Table 3. Dunn's test results. Number of variables that can differentiate between infestation levels.

Groups	Number of variables
Level 1 – level 2	1
Level 1 – level 3	30
Level 1 – Level 4	69
Level 2 – level 3	0
Level 2 – level 4	17
Level 3 – level 4	20

Table 4. Dunn's test results.

Variables	1-2	1-3	1-4	2-3	2-4	3-4
CIRE	0.8199 (NS)	0.0050 **	5.12E-10 ***	0.6725	0.0010 **	0.3294 (NS)
NDRE	0.8458 (NS)	0.0064 **	5.45E-10 ***	0.7344	0.0010 **	0.2868 (NS)
NDVI	1 (NS)	0.0392	6.05E-09 ***	1 (NS)	0.0009 ***	0.1563 (NS)
zsd	0.2923 (NS)	0.0044 **	0.0001 ***	1 (NS)	0.8502 (NS)	1 (NS)
zvar	0.2923 (NS)	0.0044 **	0.0001 ***	1 (NS)	0.8502 (NS)	1 (NS)
D1	1 (NS)	1 (NS)	0.0004 ***	1 (NS)	0.0650(NS)	0.0158 *
D0	1 (NS)	1 (NS)	0.0007 ***	1 (NS)	0.0394 *	0.0158 *
NIR mean	1 (NS)	1 (NS)	0.0035 *	1 (NS)	0.1819 (NS)	0.0048 *
Elev L2	0.5927 (NS)	0.0024 **	0.0287 *	0.6335 (NS)	1 (NS)	1 (NS)
% all returns above 3.00	1 (NS)	1 (NS)	0.0059 **	1 (NS)	0.1597 (NS)	0.1280 (NS)
D9	1 (NS)	1 (NS)	0.0373 *	1 (NS)	0.2646 (NS)	0.5458 (NS)

*: p -value < 0.05; **: p -value < 0.01; ***: p -value < 0.001; NS: non-significant.

Table 5. Support Vector Machine (Svm) and Random Forest (RF) Classification methods.

Types of metrics	Metrics	Method	Fitting phase	Validation
Spectral indices	CIRE+NDVI	SVM	0.823	0.833
	CIRE+NDVI	RF	1	0.733
Spectral indices + LiDAR	NDRE + D0	SVM	0.817	0.817
	NDVI + Elev.L2	RF	1	0.743

LiDAR	zsd + % all ret. above 3.00+ D9	SVM	0.753	0.643
	zsd + D0 + D9	RF	1	0.645

Table 6. Confusion matrix of the best SVM model with spectral indices metrics.

CIRE+NDVI	Level 1	Level 2	Level 3	Level 4	User's accuracy
Level 1	19	4	0	0	82.61%
Level 2	0	4	2	0	66.67%
Level 3	0	1	7	3	63.64%
Level 4	0	0	0	15	100.00%
Prod.'s accuracy	100.00%	44.44%	77.78%	83.33%	

Table 7. Confusion matrix of the best SVM model with LiDAR metrics.

zsd+% all ret. above 3.00+ D9	Level 1	Level 2	Level 3	Level 4	User's accuracy
Level 1	17	3	0	2	77.27%
Level 2	0	5	2	2	55.56%
Level 3	2	0	6	1	66.67%
Level 4	0	1	1	13	86.67%
Prod.'s accuracy	89.47%	55.56%	66.67%	72.22%	

Table 8. Confusion matrix of the best SVM model with spectral indices + LiDAR metrics

NDRE + D0	Level 1	Level 2	Level 3	Level 4	User's accuracy
Level 1	19	4	0	0	82.61%
Level 2	0	5	3	0	62.50%
Level 3	0	0	6	2	75.00%
Level 4	0	0	0	16	100.00%
Prod.'s accuracy	100%	55.56%	66.67%	88.89%	